



Observations of Atmospheric Methane and Carbon Dioxide Mixing Ratios: Tall-Tower or Mountain-Top Stations?

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1 **Observations of Atmospheric Methane and Carbon Dioxide**
2 **Mixing Ratios: Tall-tower or Mountain-top Stations?**

3

4 **Ines Bamberger • Brian Oney • Dominik Brunner • Stephan Henne • Markus**
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10 **Abstract** Mountain-top observations of greenhouse gas mixing ratios may be an
11 alternative to tall-tower measurements for regional scale source and sink estimation. To
12 investigate the equivalence or limitations of a mountain-top site as compared to a tall
13 tower site, we used the unique opportunity of comparing in situ measurements of
14 methane (CH₄) and carbon dioxide (CO₂) mixing ratios at a mountain top (986 m above
15 sea level, a.s.l.) with measurements from a nearby (distance 28.4 km) tall tower, sampled
16 at almost the same elevation (1009 m a.s.l.). Special attention was given to (i) how local
17 wind statistics and greenhouse gas sources and sinks at the mountain top influence the
18 observations, and (ii) whether mountain-top observations can be used as for those from a
19 tall tower for constraining regional greenhouse gas emissions. Wind statistics at the
20 mountain-top site are clearly more influenced by local flow systems than those at the tall-
21 tower site. Differences in temporal patterns of the greenhouse gas mixing ratios observed
22 at the two sites are mostly related to the influence of local sources and sinks at the

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23 mountain top site. Major influences of local sources can be removed by applying a
24 statistical filter (5th percentile) or a filter that removes periods with unfavourable flow
25 conditions. In the best case, the bias in mixing ratios between the mountain-top and the
26 tall-tower sites after the application of the wind filter was
27 -0.0005 ± 0.0010 ppm for methane (September, 0000–0400 UTC) and 0.11 ± 0.18 ppm
28 for CO₂ (February, 1200–1600 UTC). Temporal fluctuations of atmospheric CH₄ and
29 CO₂ mixing ratios at both stations also showed good agreement (apart from CO₂ during
30 summertime) as determined by moving bi-weekly Pearson correlation coefficients (up to
31 0.96 for CO₂ and 0.97 for CH₄). When only comparing mixing ratios minimally
32 influenced by local sources (low bias and high correlation coefficients), our
33 measurements indicate that mountain-top observations are comparable to tall-tower
34 observations.

35

36 **Keywords** Atmospheric observations • Greenhouse gases • Local greenhouse gas sources
37 • Mountain meteorology • Mountain top • Tall tower

38

39 **1 Introduction**

40 Tall-tower measurements of atmospheric greenhouse gases contain information from
41 larger spatial scales than measurements close to the ground (Marquis and Tans 2008) and
42 thus provide valuable information for resolving regional transport and the distribution of
43 greenhouse gases (Gloor et al., 2001). In contrast, greenhouse gas mixing ratios measured
44 at mountain-top stations are often more influenced by boundary-layer processes and by
45 local sources and sinks than are tall-tower measurements. This is mainly due to the
46 complex topography around mountain sites and the low sampling height above the
47 ground, allowing local fluxes to considerably influence the local concentration field. To
48 address the question: how similar, or how different, are mountain-top and tall-tower
49 measurements at the same elevation above sea level, we compared carbon dioxide (CO₂)
50 and methane (CH₄) mixing ratio measurements at a mountain-top and a tall-tower site
51 that are only separated by 28.4 km, which allows for a unique and direct comparison. It is
52 not expected that short-term variations at two sites separated by such a distance are
53 synchronous, but it can be expected that, with adequate statistical aggregation,

54 measurements made at both sites should provide similar regional information in the CO₂
55 and CH₄ time series.

56 While the global budgets of carbon dioxide and methane are relatively well known, their
57 sources and sinks are poorly constrained based on current regional scale observational
58 networks, where “regional” used herein refers to an area > 10⁴ km² (Cleugh et al. 2004).
59 With a joint contribution of 81 % to the positive anthropogenic radiative forcing, CO₂
60 and CH₄ are the two most important anthropogenic greenhouse gases (Myhre et al. 2013).
61 Even though tall-tower networks in both Europe and North America provide observations
62 of net changes in CO₂ and CH₄ mixing ratios at large spatial scales, the network density
63 is insufficient to resolve atmospheric and surface flux patterns at the smaller regional
64 (sub-national) scale (Marquis and Tans 2008, Villani et al. 2010, Dlugokencky et al.
65 2011). An increased number of tall-tower stations would be beneficial for resolving the
66 regional distribution of greenhouse gas mixing ratios (Lauvaux et al. 2012). A
67 concentrated effort towards a dense regional network of tall towers was initiated by the
68 NOAA Earth System Research Laboratory that resulted in a network of seven towers
69 (Andrews et al. 2014). Similarly in Europe, a tall tower network exists with seven towers
70 (e.g., Vermeulen et al. 2011, Thompson et al. 2009), and in order to expand existing and
71 create new networks, TV and radio broadcasting towers could be equipped with relevant
72 instrumentation (Marquis and Tans 2008, Andrews et al. 2014). Such towers and
73 additional sites are currently being integrated into the European Integrated Carbon
74 Observation System (ICOS, www.icos-ri.eu). However, access to existing
75 telecommunication towers may not always be possible, and the installation and
76 maintenance of measurement instrumentation is often prohibitive and time intensive.

77 Other approaches such as aircraft measurements (e.g., Beck et al. 2012, Schuck et al.
78 2012, Xiong et al. 2010) yield useful snapshots of the atmospheric composition and add
79 supplementary information on vertical and horizontal profiles (Crevoisier et al. 2006,
80 Zhang et al. 2014). Long-term aircraft measurements of greenhouse gases within the
81 NOAA atmospheric monitoring network have been carried out by Karion et al. (2013)
82 twice a week over selected areas, but such airborne measurements cannot provide
83 continuous, long-term observations, and may miss essential parts of the atmospheric
84 variability, especially on the regional scale. Remote sensing products such as satellite

85 retrievals of global atmospheric greenhouse gas total column mixing ratios have recently
86 become available (Frankenberg et al. 2005, 2008, 2015, Kuze et al. 2009, Buchwitz et al.
87 2013). These measurements are complementary in their ability to provide spatial
88 information on atmospheric greenhouse gases. However, they are still limited in their
89 temporal and spatial resolutions, are affected by cloud cover, and are less accurate and
90 less sensitive to surface fluxes than in situ measurements (Marquis and Tans 2008).
91 Alternatively, continuous measurements of CO₂ and CH₄ mixing ratios close to the
92 ground but at elevated locations such as mountain tops might provide a footprint of
93 similar size as that of a tall-tower station. Consequently, the use of such measurement
94 stations could complement the number of sites used to measure greenhouse gas mixing
95 ratios, especially in mountainous regions where tower facilities are rare (Andrews et al.
96 2014, Lee et al. 2015). However, to the best of our knowledge, the comparability of
97 mountain-top data to tall-tower data in a direct comparison, as we present here, has not
98 yet been investigated. Greenhouse gas mixing ratios at mountain tops might be
99 influenced by local flows and nearby greenhouse gas sources and sinks. All these
100 influences may negatively affect the regional representativeness of mountain top
101 measurements (Brooks et al. 2012). Before being used in atmospheric inversions,
102 mountain-top measurements may need to be screened using meteorological or statistical
103 filters in order to remove local influences (Brooks et al. 2012, Lee et al. 2015).

104 Highly accurate atmospheric mixing ratio measurements in combination with an inverse
105 transport model are required to produce quantitative information on greenhouse gas
106 sources and sinks (Nisbet and Weiss 2010). Thus, apart from network density and
107 accuracy of measurements, the top-down approach to determining regional and global
108 greenhouse gas fluxes is dependent on the reliability of transport models to allocate
109 regional sources and sinks (Lin and Gerbig 2005, Gerbig et al. 2008, Tolk et al. 2008,
110 Kretschmer et al. 2014). Even if measurements are taken at considerable altitudes above
111 ground, the exact representation of the transport of greenhouse gases may be a major
112 source of uncertainty in these top-down approaches (e.g., McKain et al. 2012, Smallman
113 et al. 2014). Especially due to the development of small-scale local circulations, such as
114 thermally induced slope-wind systems (Whiteman 2000) and their insufficient
115 representation in relatively coarse-resolution atmospheric transport models, the

116 greenhouse gas mixing ratios above complex terrain (e.g. mountain tops) are difficult to
117 predict using such models. Subgrid-scale transport processes due to complex topography
118 are currently not, or only very crudely, represented in global and regional-scale transport
119 models (van der Molen and Dolman 2007, Tolk et al. 2008) that are used for estimating
120 greenhouse gas fluxes at continental to global scales. Recently it has been argued that the
121 use of such coarse transport models results in an underestimation of net CO₂ exchange
122 (Rotach et al. 2013), which is critical since at least 50 % of the land surface can be
123 classified as complex terrain (Rotach et al. 2007, 2013). Thus, in combination with
124 measurements in complex terrain (e.g. at mountain tops), high-resolution transport
125 models are required to capture micro scale to meso scale (2 to 20 km) wind circulations
126 and their effect on greenhouse gas mixing ratios (Pillai et al. 2011). Another drawback of
127 mountain-top measurements may be that they are affected by microscale greenhouse gas
128 sources or sinks, close to the measurement inlet that cannot readily be represented in a
129 transport model. Thus, it is unclear how local sources and local flows will influence
130 greenhouse gas mixing ratios observed at a mountain-top station as compared to a tall-
131 tower station.

132 The aim of this study is to determine whether a mountain-top station can provide data of
133 similar quality as those at a tall-tower station. For this purpose, we compared CO₂ and
134 CH₄ mixing ratio measurements at sub-daily to seasonal time scales from a tall-tower and
135 a mountain-top station in Switzerland, located at a similar elevation (above sea level) and
136 with a relatively short horizontal distance from each other (28.4 km). The two sites are
137 part of a greenhouse gas observation network in Switzerland that has recently been
138 established in the framework of the CarboCount CH project (www.carbocount.ch, Oney
139 et al. 2015).

140 We pose the following questions: (1) How different are flows at the mountain-top
141 compared to the tall-tower stations? (2) How important is the potential contamination of
142 the mountain-top measurements by local greenhouse gas sources and sinks if an inlet
143 close to the ground is used? (3) Is it possible to identify optimum atmospheric conditions
144 for which mountain-top observations are comparable to those from a tall tower, given that
145 other environmental factors are similar (i.e. climate, distance to pollution from urban
146 centres)?

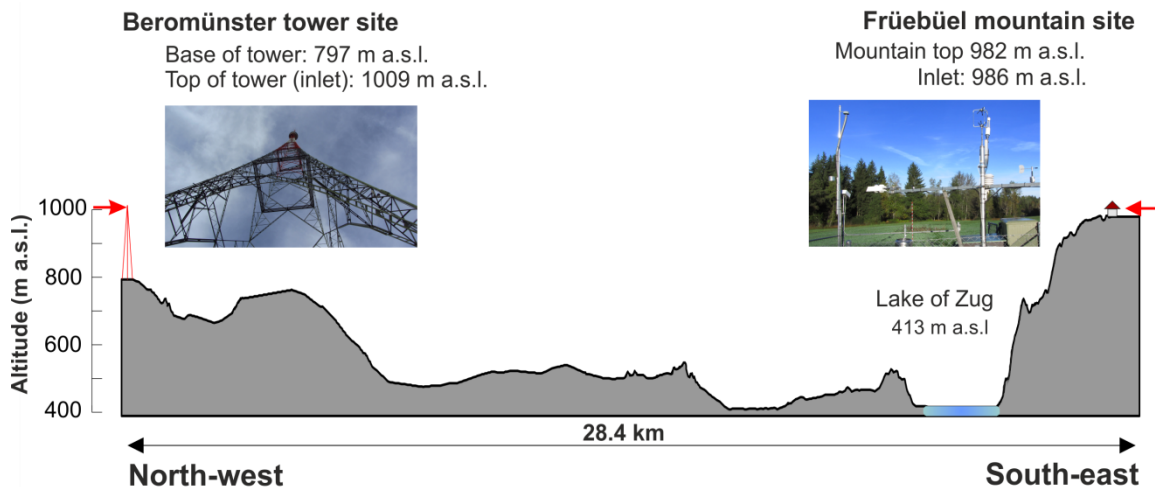
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148 2 Materials and Methods

149

150 2.1 Field Sites

151 Measurements of greenhouse gases and meteorological variables were made at two of the
152 four sites of the CarboCount CH observation network in Switzerland, which has been
153 established to investigate greenhouse gas sources and sinks through an integrated
154 approach using both observations and atmospheric transport models (Oney et al. 2015).
155 The two sites are located within 28.4 km of each other close to the southern edge of the
156 Swiss Plateau, the relatively flat, most densely populated and agriculturally used area in
157 Switzerland, located between the Alps and the Jura mountains. A cross-sectional profile
158 of the topography between both stations and photographs of the corresponding sites are
159 shown in Fig. 1. The proximity of both sites guaranteed comparable weather conditions
160 as well as a similar distance to major greenhouse gas sources and sinks.



161

162 **Fig. 1** Topographical transect between the tall-tower site Beromünster (left) and the mountain-top site
163 Fruebüel (right). Red arrows indicate the inlet heights at both sites. Note the vertical amplification of the
164 scale.

165

166 2.1.1 Fruebüel Mountain-top Station

167 Measurements at Fruebüel – a Swiss FluxNet site (www.swissfluxnet.ch) – were taken on
168 top of the Zugerberg mountain ridge, located at an elevation of 982 m above sea level

169 (a.s.l.) in the pre-alpine foothills of Switzerland (47°06'57.0" N, 8°32'16.0" E). The site is
170 located on a relatively flat highland, extending about 2 km in the east–west and 5 km in
171 the north–south directions, and is thus not a classical mountain-top site because it has no
172 pronounced peak. Average daytime convective boundary-layer heights above the Swiss
173 Plateau usually exceed 500 m above the ground during the warm season (March to
174 September, Collaud Coen et al. 2014). With an elevation of 1000 m (\approx 500 m above the
175 Swiss Plateau) we therefore assume that Frübüel mostly lies within the convective
176 boundary layer.

177 The Zugerberg mountain ridge is a sparsely populated agricultural area, with prevailing
178 land-cover types managed grasslands and forests. The towns of Zug (at \approx 10 km distance)
179 and Lucerne (\approx 20 km) are the nearest larger towns and are situated approximately
180 450 m below the measurement station's elevation.

181 The air inlet for the greenhouse gas mixing ratio measurements was installed 4 m above a
182 moderately intensively managed grassland (Zeeman et al. 2010, Imer et al. 2013), with
183 meteorological measurements (wind speed, air temperature) taken approximately 3 m
184 away from the air inlet at 2-m height. In 2013, the average air temperature was 7.1 °C,
185 ranging from -14.5 °C to 34.8 °C (range of 2-h averages).

186 Farmsteads and barns of the ETH (*Eidgenössische Technische Hochschule*) research
187 station Frübüel are located approximately 300 m to the south-west of the atmospheric
188 measurement tower. In 2013, the farmstead accommodated between 96 (June) and 50
189 (December) head of cattle, and from October till mid-November an additional 15 sheep.
190 From mid-July to mid-September all cattle were relocated to remote alpine pastures off
191 site. Otherwise, there were no other farmsteads or other anthropogenic sources in close
192 vicinity to the site. The three grassland parcels directly adjacent to the station were
193 managed differently; while the ungrazed meadow next to the measurement station was
194 cut three times and fertilized twice a year, the two grazed parcels were only fertilized
195 once and cut once or twice a year (Table S1 in Supplementary Material).

196

197 *2.1.2 Beromünster Tall-Tower Station*

198 The Beromünster tall-tower station is located approximately 30 km to the west of
199 Frübüel, atop a gentle hill (797 m a.s.l., 47°11'22.4"N, 8°10'31.5"E) close to Lake
200 Sempach, Switzerland. The area surrounding the tall-tower station is predominantly used
201 for agriculture. The two larger towns closest to the tall tower are Lucerne (≈ 20 km) and
202 Zug (≈ 30 km). The tall tower is the taller of two former national radio broadcast towers
203 and is protected as a national monument, and is no longer used for telecommunication.
204 Although greenhouse gas and meteorological measurements were made at five different
205 heights, we have only used measurements from the top of the tower at an elevation of
206 1009 m a.s.l. (212 m above ground level), similar to the elevation of measurements at
207 Frübüel. Air temperatures recorded at this elevation ranged from -9.9 °C to 29.1 °C in
208 the year 2013, resulting in an annual average of 7.3 °C.

209

210 *2.2 Instrumentation*

211 Mixing ratios of CO₂ and CH₄ at the mountain-top station (Frübüel) were measured with
212 a cavity ring-down spectrometer (Picarro G2301, Picarro Inc., Santa Clara, California,
213 USA), while the tall-tower station (Beromünster) was equipped with a Picarro G2401
214 analyzer that also measured CO mixing ratios. Both instruments were operated at a
215 standard cavity pressure of 186.7 hPa and a cavity temperature of 45 °C. All times
216 reported here are given in UTC, which differs by one hour from local time (CET-1 h).

217 **Frübüel:** The air inlet was covered by an inlet filter (F-15-050, Solberg International,
218 Ltd., Itasca, IL, USA). Ambient air was drawn through an insulated and heated Synflex
219 1300 hose to the analyzer, which was placed within an air-conditioned (20–28 °C)
220 container. Automatic recalibration of the system was performed daily using three
221 reference gases (high: CO₂/CH₄ mixing ratios 475.3/2.398 ppm; low: 383.5/1.895 ppm;
222 working: 405.2/2.056 ppm). Calibration gases were traced to the WMO primary
223 standards WMO-X2007 for CO₂ and WMO-X2004 for CH₄. The accuracy of the
224 greenhouse gas measurements was estimated at ≤ 0.07 ppm and ≤ 0.0005 ppm for CO₂
225 and CH₄, respectively (Oney et al. 2015). Details about the meteorological sensors are
226 given by Zeeman et al. (2010).

227 **Beromünster:** At the Beromünster tall tower, air was drawn at a flow rate of 14 L min^{-1}
228 from the highest level (212 m) to the analyzer through a 220 m long inlet tube. The
229 analyzer was placed within the old radio transmitter building at the base of the tower, and
230 automatic calibration was done once a week (high: 472.66/2.4247 ppm; low:
231 382.11/1.9089 ppm) similarly to Frübüel. In addition, every 6 h measurements were
232 compared with a working gas standard (392.24/2.1312 ppm), see Berhanu et al. (2016).

233 Meteorological measurements at the highest level of the tall tower were provided by an
234 integrated weather station (MetPakTM II Remote, Gill Instruments Ltd, Lymington,
235 Hampshire, UK). The weather station included sensors to measure wind direction and
236 speed, air temperature, relative humidity, and barometric pressure. All measurements
237 were performed at 1-s resolution and processed to 2-h averages for data analysis.

238

239 2.3 Data Analysis

240 Comparisons between the two sites initially used 1-min average mixing ratios aggregated
241 to 2-h intervals if at least 20 % of the 1-min averages represented concurrent
242 measurements at both sites. For each 2-h interval the following two statistical parameters
243 were computed: (1) the average greenhouse gas mixing ratios; and (2) the quantile of the
244 CO_2 and CH_4 frequency distribution within each 2-h interval based on 1-min data, which
245 separates the lowest 5 % from the upper 95 % of the observed mixing ratios (referred to
246 as the 5th percentile). Similar to the arithmetic mean or the median (= 50th percentile), the
247 5th percentile acts as a low-pass filter. This was necessary since the frequency distribution
248 of CH_4 mixing ratios is positively skewed towards higher mixing ratios at the mountain-
249 top station. In contrast, the lower tail of the frequency distribution is much less influenced
250 by outliers, thus suggesting that the 5th percentile is more likely to represent air parcels
251 unaffected by local sources than the arithmetic mean. One drawback of the 5th percentile,
252 however, is its high sensitivity to local sinks (for example the biospheric sink for CO_2).
253 Two-hourly averages and 5th percentile values were used for further analysis of the
254 atmospheric greenhouse gas mixing ratios at the two stations.

255 To identify a filter criterion for the mountain-top station that eliminates unwanted local
256 influences that do not relate to regional-scale greenhouse gas fluxes, we evaluated

257 differences between mountain-top and tall-tower mixing ratios using these aggregated
258 data. We calculated binned average differences according to wind speed, wind direction
259 and atmospheric stability at the mountain-top site (see Supplementary Material, Section
260 S2).

261 To determine the level of agreement between wind-direction measurements at the two
262 sites, we used a kernel distribution in a polar coordinate system. The kernel density
263 estimation of wind direction was calculated using the procedure described by Botev et al.
264 (2010). As differences in CH₄ mixing ratios between the mountain-top and the tall-tower
265 sites may be due to local CH₄ sources at the mountain-top site, we investigated between-
266 site differences in CH₄ mixing ratios and their dependence on wind direction observed at
267 the mountain-top site. To this end, we grouped the differences between the corresponding
268 2-h tall-tower and mountain-top mixing ratios into 18 wind sectors (each spanning 20 °).
269 Then, the average differences in CH₄ mixing ratios between the sites were calculated for
270 daytime (0800–1600 UTC) and nighttime (2000–0400 UTC), summer (April to October)
271 and winter (November to March) and for each wind sector. More detailed information on
272 the aggregation of these classes is presented in the Supplementary Material (Section S3,
273 Figs. S2 and S4). For CO₂, the same procedure was repeated using a slightly modified
274 temporal (morning: 0800–1200 UTC, afternoon: 1200–1600 UTC, evening: 2000–2400
275 UTC and night: 0000–0400 UTC) and seasonal classification (spring: April to June,
276 summer/autumn: July to October, winter: November to March) to account for the
277 differences in sources and sinks of CO₂ in comparison with CH₄.

278 In order to determine the general agreement between the two sites, we used the 5th
279 percentile time series and the time series after removing periods with unfavourable flow
280 conditions for CO₂ and CH₄ and calculated the bias and Pearson's correlation coefficients
281 between the tall-tower and the mountain-top stations. The bias after filtering was
282 evaluated for different months and times of day, whereas the correlation coefficient was
283 calculated between the filtered time series of CO₂ and CH₄ mixing ratios within time
284 periods of ± 7 days centred at each 2-h value.

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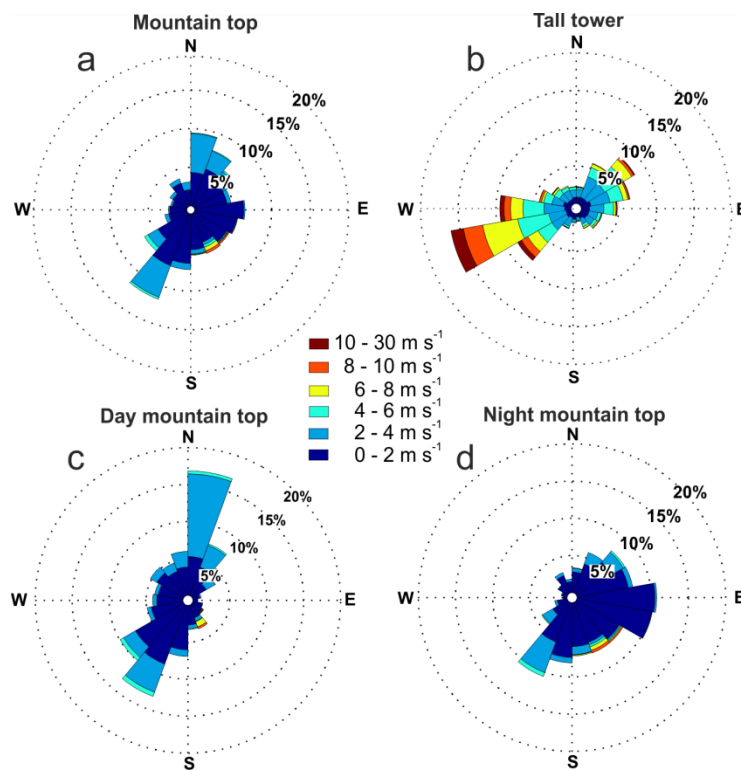
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287 **3 Results**

288

289 3.1 Wind Directions at the Tall-tower and the Mountain-top Site

290 Wind directions at the mountain-top site were similar to those at the tall-tower site (Fig.
291 2a, b), but with a few noteworthy differences. Wind directions at the tall-tower site were
292 dominated by the typical south-westerly and north-easterly directions found over the
293 Swiss Plateau due to the channelling of flow between the Alps and the Jura mountains
294 (Wanner and Furger 1990). In contrast, the flow at the mountain-top site was generally
295 much weaker and arriving from more variable wind directions. Strongest wind at the
296 mountain-top site was observed from the south-easterly sector (around 150 °), most
297 probably corresponding to foehn wind.



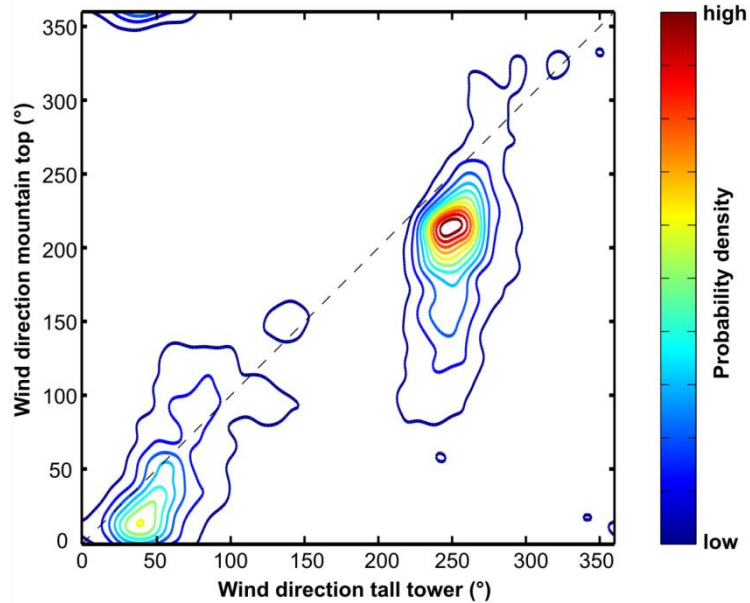
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299 **Fig. 2** Wind-rose plots for the measurement stations Frübüel (mountain top, **a**) and Beromünster (tall
300 tower, **b**) during the year 2013 additionally separated for daytime (0800–1600 UTC, **c**) and nighttime
301 (0000–0400 UTC, **d**) at the mountain-top station

302

303 To emphasize the influence of diurnal variations due to thermally-forced slope flow, Fig.
304 2c, d show wind roses at the mountain-top site separately for daytime and nighttime.

305 While flow from northerly directions was most frequent during daytime, nighttime flow
306 was predominantly from southerly or easterly directions. Flow from easterly and westerly
307 directions was generally weak, more frequently arriving from the east at night and from
308 the west during the day.



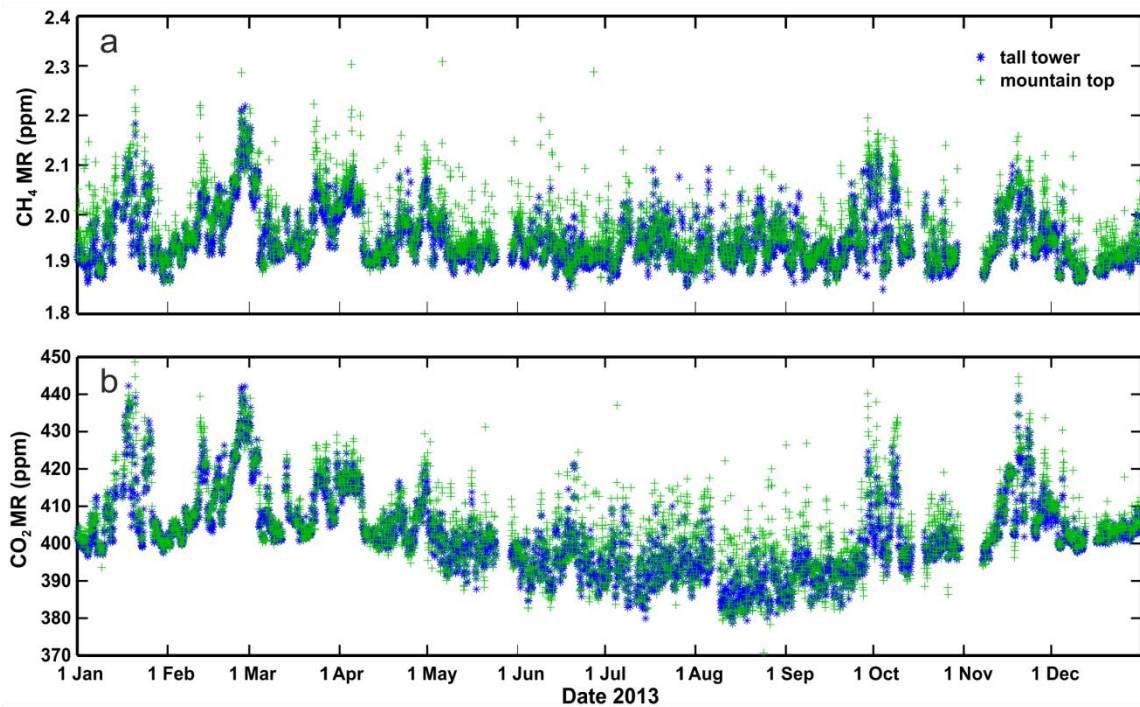
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310 **Fig. 3** Probability density distribution of wind directions at the Beromünster tall-tower station (horizontal
311 axis) versus the Früebüel mountain-top station (vertical axis) based on a bivariate kernel density estimation.
312 Reddish colours denote a high probability density, while bluish colours depict lower probability densities

313
314 The kernel density plot of the bivariate probability distribution between wind directions
315 (Fig. 3) showed that flows from the north-easterly sector at the tall-tower site were
316 accompanied by flows from a similar direction also at the mountain-top site. Larger
317 directional spread was found at the mountain-top site, whereas the tall-tower site
318 experienced rather well defined south-westerly flow (220 °–270 °). Under these
319 conditions, the wind directions at Früebüel varied over a wider range of angles, but the
320 peak of the joint probability distribution showed a relatively small directional bias
321 ($\approx 20^\circ$ – 30°), likely reflecting the influence of the surface drag at the mountain top
322 compared to the tower measurements that are more representative of mid-boundary layer
323 conditions.

324

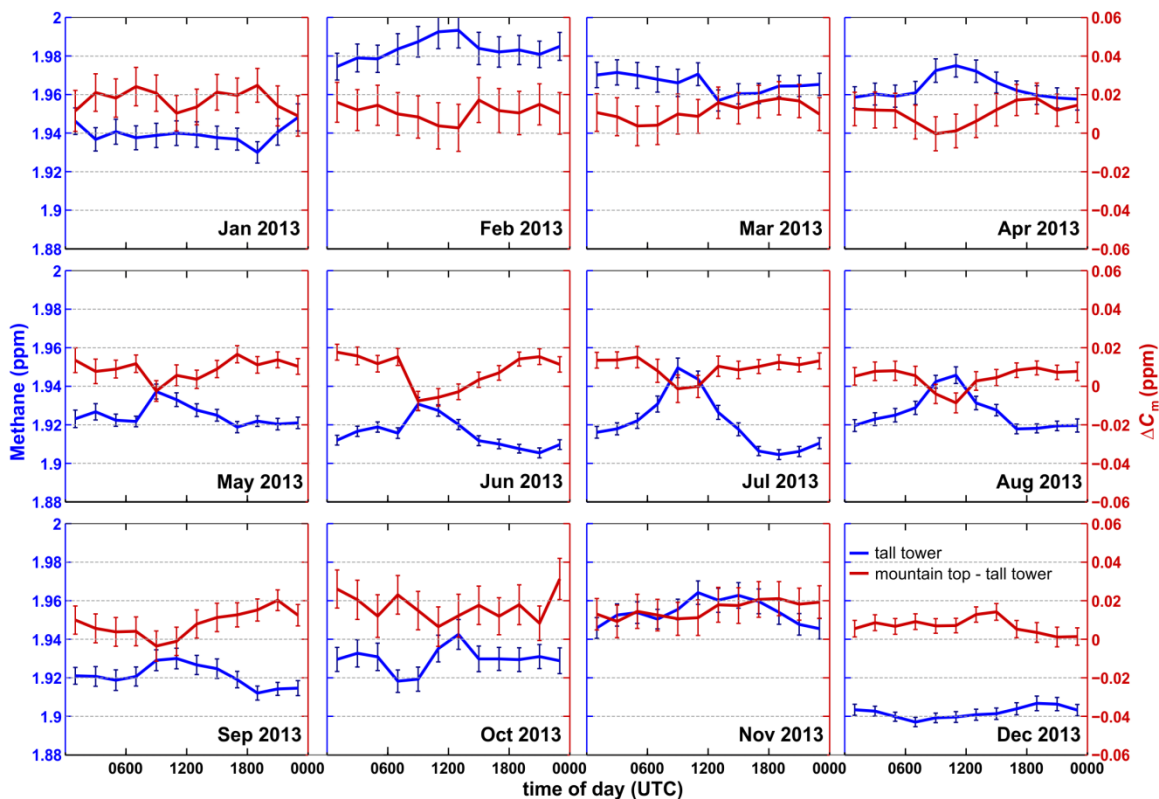
325 3.2 Greenhouse Gas Variations

326 **Methane:** Despite notable differences in wind direction and wind speed between the two
327 sites (Figs. 2 and 3), CH₄ mixing ratios followed a similar seasonal course (Fig. 4a), with
328 mixing ratios varying roughly between 1.8 ppm and 2.2 ppm. Peak CH₄ mixing ratios
329 observed at the mountain-top site, however, deviated by up to 0.2 ppm from mixing ratios
330 observed at the tall-tower site. The positive outliers were most likely influenced by local
331 CH₄ sources at the mountain-top site.



332
333 **Fig. 4** Time series of 2-h averages of CH₄ (a), and CO₂ (b) mixing ratios measured at the Beromünster tall-
334 tower station (blue symbols) compared to those measured at the Früebüel mountain-top station (green
335 symbols) from January to December 2013

336
337 For average CH₄ mixing ratios these differences between the two stations remained
338 generally below 0.02 ppm for any wind direction, except for southerly flow from the
339 direction of the ETH research station when average mixing ratios at the mountain-top site
340 were considerably higher (up to 0.065 ppm, Fig. S1 in Supplementary Material). This
341 pattern indicates that the major source for methane at this site is the farmstead of the ETH
342 research station.



343

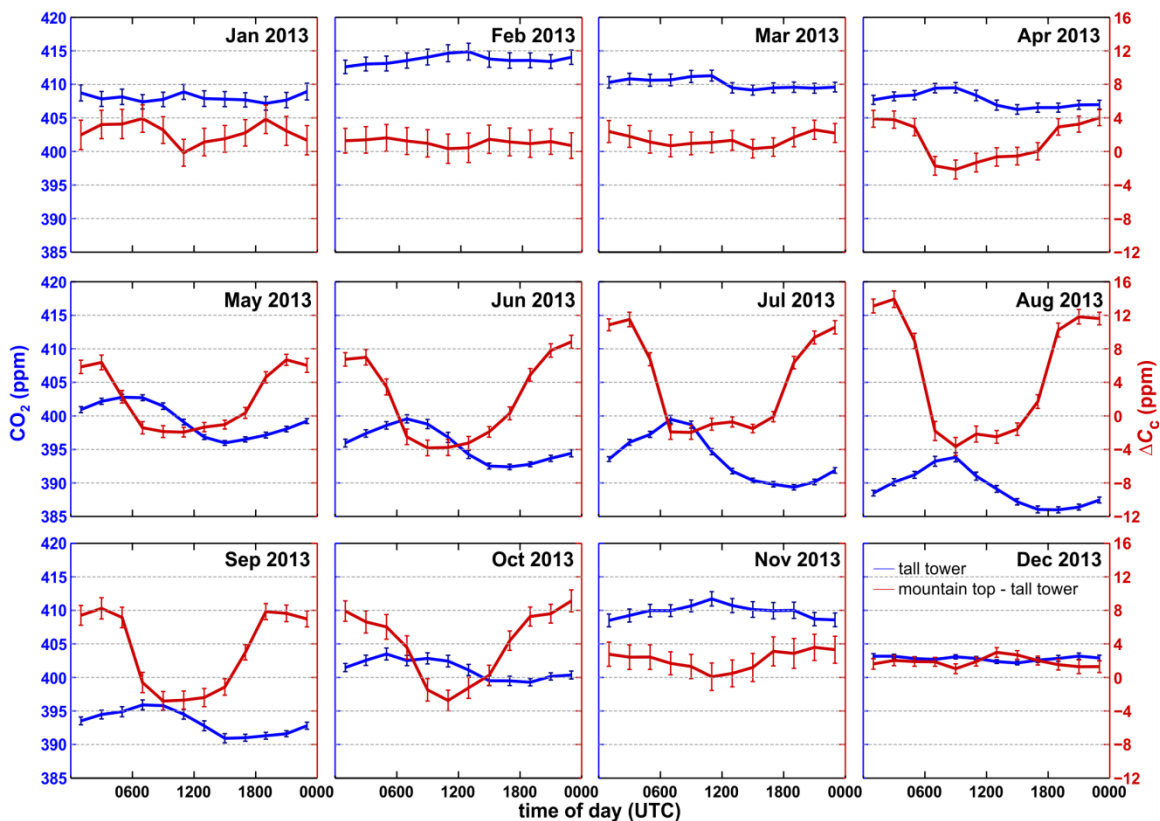
344 **Fig. 5** Monthly average diurnal cycles derived from the 5th percentiles of the 2-h frequency distribution of
 345 CH₄ mixing ratios observed at the Beromünster tall-tower station (blue colour) and the difference in CH₄
 346 mixing ratios between the mountain-top and the tall-tower stations (ΔC_m , red colour) including the
 347 corresponding 95% confidence interval (vertical bars) from January to December 2013

348

349 The influence of local sources was investigated by analyzing monthly aggregated diurnal
 350 patterns of the 5th percentiles of the CH₄ mixing ratios at 2-h resolution. Such CH₄
 351 mixing ratios only showed weak diurnal patterns at the mountain top, whereas an increase
 352 of up to 0.025 ppm was seen in the late morning or around midday at the tall-tower
 353 station, especially during the summer months (Fig. 5). Except for the late
 354 morning/midday minima when methane mixing ratios at the mountain-top station were
 355 lower than at the tall-tower station, monthly averages of CH₄ mixing ratios were
 356 moderately but significantly higher at the mountain-top site (between 0.005 ppm in
 357 August and 0.020 ppm in January) than at the tall-tower site (Fig. 5, red line). A generally
 358 weak bimodal seasonal pattern of methane mixing ratios was observed. At both sites,
 359 mixing ratios were low during winter (December and January), reached their first

360 maximum during February, then declined slowly during the summer and exhibited a
361 second maximum in November.

362 **Carbon dioxide:** During winter, CO₂ mixing ratios at the mountain-top and tall-tower
363 stations almost followed the same temporal course, mostly fluctuating between 400 ppm
364 (daytime) and 440 ppm (nighttime; Fig. 4b). In summer, however, the average CO₂
365 mixing ratios at the mountain top station were lower and varied between 380 ppm and
366 420 ppm. In general, CO₂ mixing ratios in summer were lower than in winter.



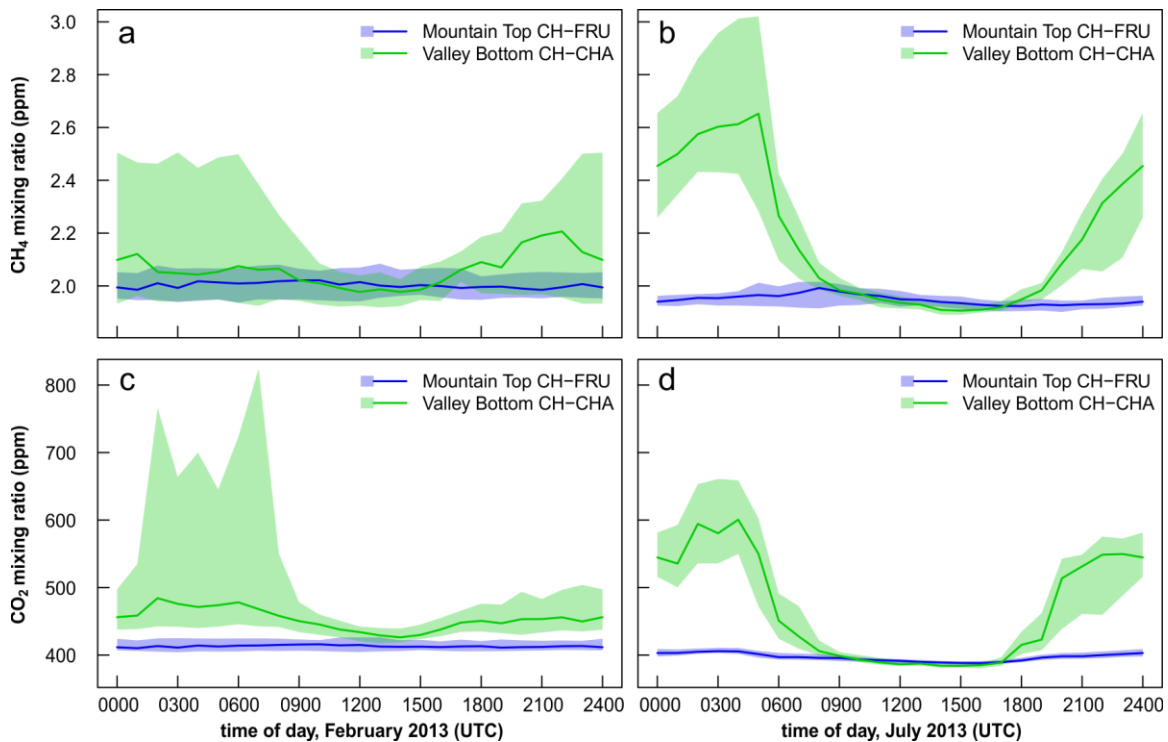
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368 **Fig. 6** Monthly average diurnal cycles of the 2-h frequency distribution of CO₂ mixing ratios observed at
369 the Beromünster tall-tower station (blue colour) and the difference in CO₂ mixing ratios between the tall-
370 tower and the mountain-top stations (ΔC_c , red colour) including the corresponding 95% confidence interval
371 (vertical bars) from January to December 2013

372

373 To investigate the difference in summertime CO₂ mixing ratios between the two sites,
374 average diurnal cycles of CO₂ for each month and both stations were calculated in an
375 analogous manner to CH₄, but using averages instead of 5th percentiles (since the 5th

376 percentile approach did not work satisfactorily for CO₂) (Fig. 6). During wintertime, only
 377 a weak diurnal cycle of CO₂ mixing ratios could be seen at the tall-tower station, showing
 378 a slight increase of up to 4 ppm during midday. This increase in midday CO₂ mixing
 379 ratios was not observed at the mountain-top station and thus the bias between the
 380 mountain-top and the tall-tower stations (red line in Fig. 6) usually decreased around this
 381 time of day. CO₂ mixing ratios always remained close to 410 ppm (except during
 382 December when CO₂ mixing ratios approached 400 ppm). In summer, CO₂ mixing ratios
 383 at the mountain-top site showed a typical diurnal cycle with lowest values during daytime
 384 and a maximum at night. Summertime CO₂ mixing ratios at the tall-tower site had their
 385 maximum in the morning and reached lowest values in the afternoon. Thus, the bias in
 386 CO₂ mixing ratios was usually negative during the day and positive at nighttime.



387

388 **Fig. 7** Diurnal cycles of median CH₄ (a, b) and CO₂ (c, d) mixing ratios at the mountain-top station
 389 (Früebüel CH-FRU, blue colour) and a nearby valley-bottom station at 393 m a.s.l. (Chamau CH-CHA
 390 47°12'37"N, 8°24'38"E, green colour; Merbold et al. 2014) for a representative month in the winter
 391 (February 2013, a, c) and summer (July 2013, b, d). The green and blue shaded areas indicate the
 392 interquartile range for the valley bottom and mountain-top station, respectively

393

394 Near-surface CH₄ mixing ratios at the mountain-top site showed a less distinct diurnal
395 cycle than at a nearby (ca. 15 km distance) valley-bottom site (Chamau, cf. Fig.7), where
396 nighttime CH₄ mixing ratios increased up to 2.4 ppm and 2.2 ppm in July and February,
397 respectively. These valley-bottom data were collected and published by Merbold et al.
398 (2014) and are only shown for comparison. A similar difference between mountain-top
399 and valley-bottom sites is seen in CO₂ mixing ratios, with a very weak nighttime increase
400 of approximately 20 ppm at the mountain-top site in July 2013 as compared to the 200-
401 ppm increase observed at the valley-bottom site.

402

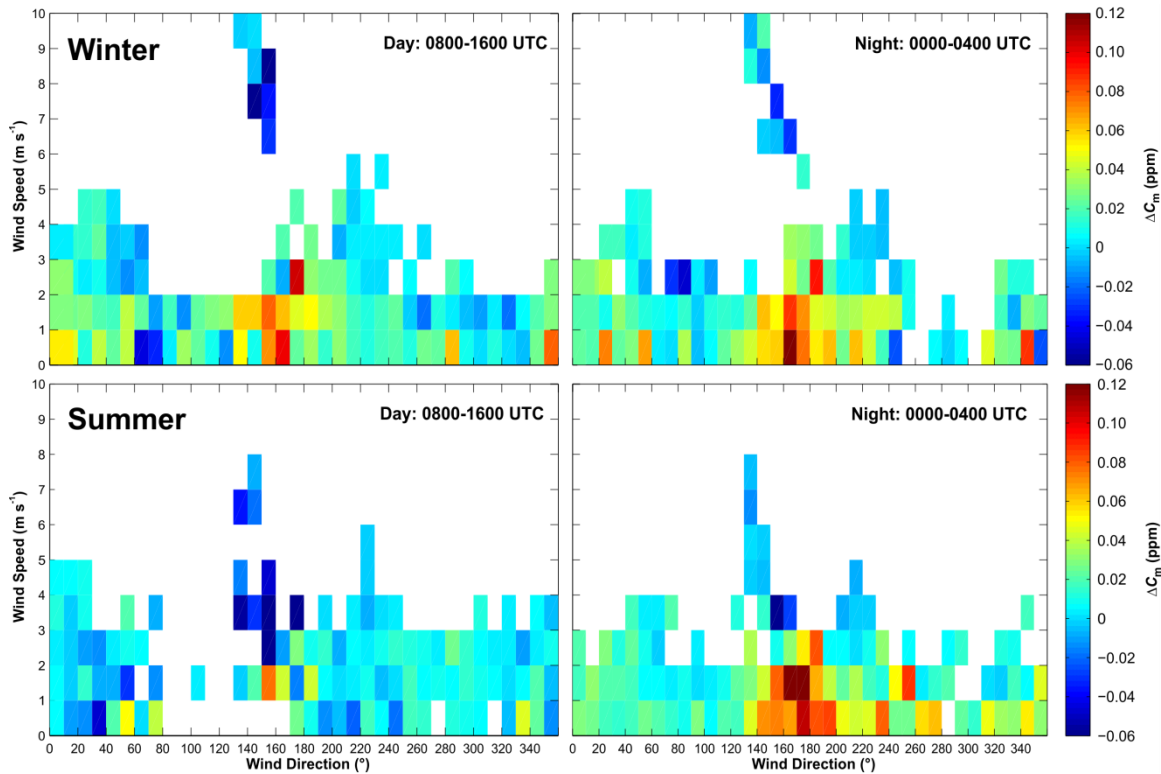
403 3.3. Filtering CH₄ and CO₂ and Mixing Ratio Differences to Minimize Local Influences

404 The percentile filtering approach revealed to be highly capable of removing a relevant
405 share of mixing ratios influenced by local sources, but is not based on physical
406 considerations, since it only uses a statistical filtering criterion. Hence, we also applied a
407 second filtering approach that employs meteorological information. With this second
408 approach, wind speeds and wind directions could be identified that led to large
409 differences between mountain-top and tall-tower measurements. These were then
410 discarded. From the retained subset of measurements, the signal with minimal influence
411 of local conditions was extracted as a function of time of day and season. An analysis of
412 variance approach was used to determine the aggregation levels for wind speed, wind
413 direction, time of day, and seasonality. The aggregation levels differed slightly for CH₄
414 and CO₂ mixing ratios. Details are given in Supplementary Material, Sections S3.1 and
415 S3.2. For simplicity, we termed this meteorological filtering the “WDS filter” (W for
416 wind speed and direction, D for time of day, and S for seasonality).

417 The differences of CH₄ mixing ratios between the mountain-top and the tall-tower sites
418 (ΔC_m) were best explained by the four factors: wind direction (three classes), wind speed
419 (five classes), time of day (separated into daytime 0800–1600 UTC and nighttime 0000–
420 0400 UTC), and seasonality (two classes: summer, April to October; and winter,
421 November to March). For CO₂, the between-site difference of mixing ratios (ΔC_c)
422 depended on the same four factors but required a higher degree of detail, with four
423 classes for time of day (morning: 0800–1200 UTC, afternoon: 1200–1600, evening:

424 2000–2400 UTC, and night: 0000–0400 UTC) and three classes for season (spring: April
 425 to June, summer/autumn: July to October, and winter: November to March). Transition
 426 periods between night and morning (0400–0800 UTC) and between afternoon and
 427 evening (1600–2000 UTC) were excluded from this analysis since during these time
 428 periods conditions are changing rapidly and may result in mixing ratios that are
 429 representative of neither daytime nor of nighttime conditions. This classification was
 430 coarse enough to be meaningful but still detailed enough to represent the pronounced
 431 diurnal cycle of the CO₂ mixing ratios (Fig. 6).

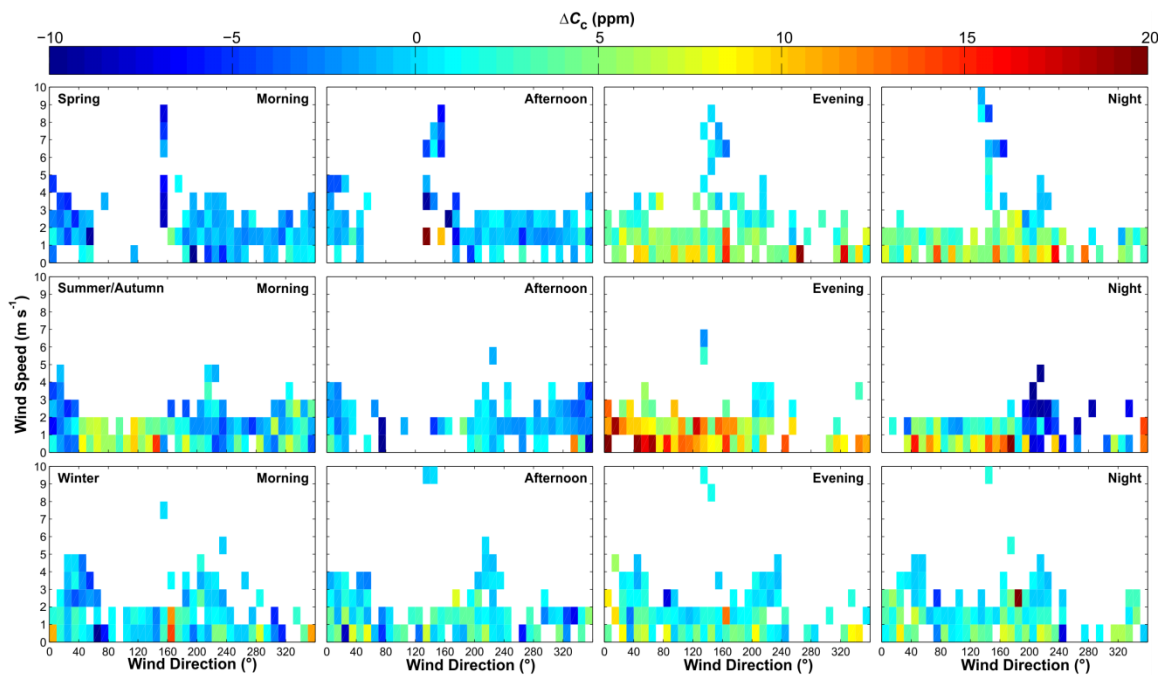
432 Before applying such a filter, a clear dependence of ΔC_m and ΔC_c mixing ratio
 433 differences on all four factors was obvious (Figs. 8 and 9). Wind direction had a
 434 comparatively larger influence on ΔC_m (Fig. 8) than ΔC_c (Fig. 9).



435
 436 **Fig. 8** Average differences of daytime (0800–1600 UTC, left) and nighttime (0000–0400 UTC, right) CH₄
 437 mixing ratios between mountain-top and tall-tower station (ΔC_m) during the winter season (November–
 438 March, top) and the summer season (April–October, bottom), binned for different wind direction and wind
 439 speed classes at the mountain top

440

441 Under strong ($> 3 \text{ m s}^{-1}$) south-easterly flow (130° to 170°), CH_4 mixing ratios tended to
 442 be up to 0.06 ppm lower at the mountain-top site compared to the tall-tower site, especially during daytime (Fig. 8).
 443 In contrast, CH_4 mixing ratios were up to 0.07 ppm higher at the mountain-top site compared to the tall-tower site with wind speeds $< 3 \text{ m s}^{-1}$
 444 and wind directions from 120° to 200° corresponding to the sectors influenced by the
 445 ETH farmstead (Fig. S1). Negative ΔC_m mixing ratio differences were also observed
 446 during easterly flow (60° – 120°) with wind speeds $> 2 \text{ m s}^{-1}$. Especially in summer,
 447 positive deviations of the mountain-top from the tall-tower measurements tended to be
 448 higher during the night than during the day.
 449



450

451 **Fig. 9** Average differences of morning (0800–1200 UTC, left), afternoon (1200–1600 UTC), evening
 452 (2000–2400 UTC), and night (0000–0400 UTC) CO_2 mixing ratios between mountain-top and tall-tower
 453 station (ΔC_c) during spring (April–June, top), summer/autumn (July–October, centre) and the winter season
 454 (November–March, bottom) binned for different wind direction and wind speed classes at the mountain top

455

456 The weak wind direction dependency of ΔC_c (Fig. 9) compared to ΔC_m (Fig. 8) is not
 457 surprising because the main local source and sink for CO_2 at the mountain top is the
 458 vegetation which extends around the measurement station in all directions. During spring,
 459 daytime CO_2 mixing ratios (morning, afternoon) at the mountain-top station were

460 generally lower than at the tall-tower station, irrespective of wind direction. Contrastingly,
461 CO₂ mixing ratios measured during evening and night were increased at the mountain-top
462 station as compared to the tall-tower station. Similar conditions were found in
463 summer/autumn except with easterly wind directions in the morning ($\Delta C_c > 0$ ppm) and
464 with south-westerly wind directions during the night ($\Delta C_c < 0$ ppm). In winter, ΔC_c was
465 generally smaller than in summer, and wind directions 90 °–160 ° and 190 °–260 ° tended
466 to show the smallest differences between the mountain-top and the tall-tower
467 measurements (Fig. 9).

468

469 3.4 Comparison of Greenhouse Gas Mixing Ratios

470 All wind directions with strong local influence on CH₄ or CO₂ mixing ratios were filtered
471 out based on the analysis of variance results shown in Supplementary Material (Tables S2
472 and S3). CH₄ mixing ratios were maintained if the following conditions were met: (1)
473 summer daytime wind directions in the range 0 °–50 ° or 190 °–360 °, (2) summer
474 nighttime wind directions in the range 10 °–120 ° or 200 °–360 ° in combination with
475 wind speeds $> 2 \text{ m s}^{-1}$, (3) winter daytime wind directions in the range 0 °–70 °
476 combined with wind speeds $> 2 \text{ m s}^{-1}$, (4) winter daytime wind directions in the range
477 220 °–360 ° combined with wind speeds $> 1 \text{ m s}^{-1}$, (5) winter nighttime wind speeds in
478 the range 1–2 m s^{-1} and wind directions in the range 40 °–130 °, and (6) winter nighttime
479 wind directions in the range 200 °–260 ° and wind speeds $> 2 \text{ m s}^{-1}$. The remaining CH₄
480 mixing ratios were used for further analysis.

481 CO₂ mixing ratios at the mountain-top station during spring or autumn were generally
482 either higher or lower than at the tall-tower station, thus filtering of CO₂ mixing ratios
483 was problematic: In winter only CO₂ mixing ratios could be retained when wind direction
484 was 90 °–160 ° or 190 °–260 ° in combination with wind speeds $> 1 \text{ m s}^{-1}$ irrespective of
485 time of day. In summer, all CO₂ mixing ratios at the mountain-top station were strongly
486 affected by CO₂ uptake (photosynthesis) by vegetation during daytime and CO₂ release
487 (respiration) during the night. With this data screening, 35 % of the CH₄ mixing ratios
488 and 9 % of the CO₂ mixing ratios measured at the mountain-top station could be retained
489 (Fig. 10 a, c).

490

491 **Table 1** Average differences between the CH₄ (ΔC_m) and CO₂ (ΔC_c) mixing ratios at the
 492 mountain-top and the tall-tower station after the application of the WDS filter

average ΔC_m mixing ratio difference \pm standard deviation ΔC_m (ppm)				
	Night	Morning	Afternoon	Evening
January	-0.0018 ± 0.0026	0.0392 ± 0.0041	0.0252 ± 0.0025	0.0030 ± 0.0017
February	0.0041 ± 0.0009	0.0008 ± 0.0024	0.0038 ± 0.0023	0.0121 ± 0.0016
March	0.0077 ± 0.0015	0.0030 ± 0.0033	0.0135 ± 0.0018	0.0144 ± 0.0019
April	0.0074 ± 0.0013	-0.0050 ± 0.0018	0.0084 ± 0.0011	0.0167 ± 0.0016
May	0.0083 ± 0.0013	0.0082 ± 0.0016	0.0157 ± 0.0013	0.0296 ± 0.0028
June	0.0319 ± 0.0030	-0.0019 ± 0.0018	0.0043 ± 0.0013	0.0348 ± 0.0035
July	0.0091 ± 0.0018	-0.0052 ± 0.0026	0.0113 ± 0.0017	0.0071 ± 0.0010
August	-0.0018 ± 0.0010	-0.0149 ± 0.0022	0.0033 ± 0.0019	0.0047 ± 0.0012
September	-0.0005 ± 0.0010	-0.0080 ± 0.0021	0.0115 ± 0.0030	0.0156 ± 0.0028
October	0.0110 ± 0.0038	0.0141 ± 0.0034	0.0151 ± 0.0031	0.0262 ± 0.0044
November	0.0037 ± 0.0008	-0.0040 ± 0.0025	0.0145 ± 0.0016	0.0157 ± 0.0020
December	0.0106 ± 0.0022	0.0031 ± 0.0020	0.0079 ± 0.0019	-0.0007 ± 0.0014
average ΔC_c mixing ratio difference \pm standard deviation ΔC_c (ppm)				
	Night	Morning	Afternoon	Evening
January	1.01 ± 0.27	0.26 ± 0.18	1.99 ± 0.34	0.91 ± 0.25
February	-0.71 ± 0.21	-0.19 ± 0.19	0.11 ± 0.18	-0.27 ± 0.18
March	1.00 ± 0.18	2.84 ± 0.60	2.82 ± 0.73	2.35 ± 0.36
November	1.76 ± 0.26	1.69 ± 0.45	2.19 ± 0.64	1.34 ± 0.22
December	1.32 ± 0.24	1.05 ± 0.15	2.41 ± 0.26	0.65 ± 0.13

493 ΔC_m , ΔC_c : Difference in CH₄ and CO₂ mixing ratios between the mountain-top and the tall-tower station
 494 derived as average value of each 2-h bias passing the WDS filter within the corresponding time period
 495 (including the corresponding standard deviation). Night: 0000-0400 UTC; morning: 0800–1200 UTC;
 496 afternoon: 1200–1600 UTC; evening: 2000–2400 UTC. The WDS filter is a meteorological conditional
 497 filter described in Section 3.3.

498

499 These retained data are thought to best represent the regional atmospheric signal in a way
 500 that allows for a direct comparison of the mountain-top mixing ratios with those
 501 concurrently measured at the tall-tower station. Average ΔC_m after WDS filtering
 502 (Table 1) varied between -0.0149 ± 0.0022 ppm (0800–1200 UTC in August) and
 503 0.0392 ± 0.0041 ppm (0800–1200 UTC in January). In September between 0000 and
 504 0400 UTC, the average ΔC_m bias was closest to zero (-0.0005 ± 0.0010 ppm). During the

505 morning hours, the average ΔC_m was often negative (tall-tower station observed higher
506 values than the mountain-top station), especially in summer. Average ΔC_c after filtering
507 varied between -0.71 ± 0.21 ppm (0000–0400 UTC in February) and 2.84 ± 0.60 ppm
508 (0800–1200 UTC in March) and was 0.11 ± 0.18 ppm (1200–1600 UTC in February) in
509 the best case.

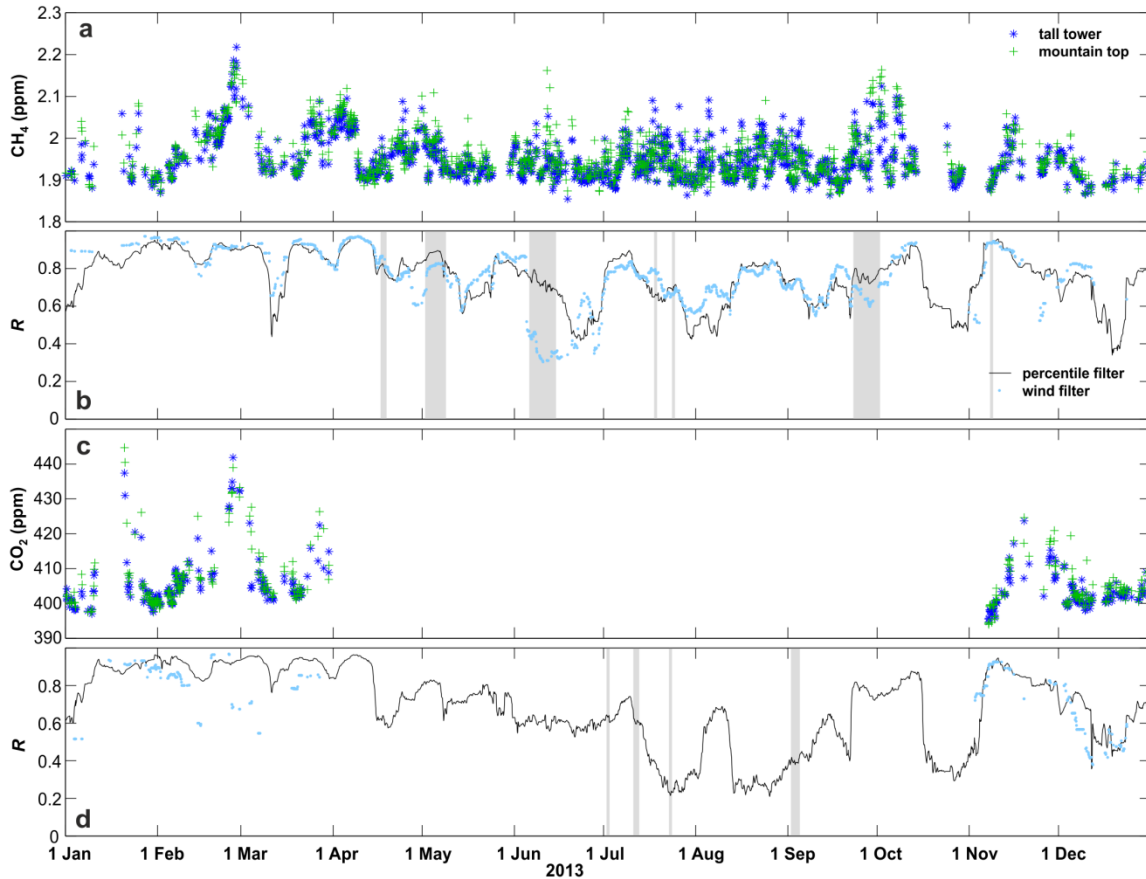
510 In addition to the bias calculations we also performed a correlation analysis with (a) the
511 WDS filtered 2-h mixing ratio averages and (b) the 5th percentiles of CH₄ and CO₂
512 mixing ratios measured at both sites. Fourteen-day running Pearson's correlation
513 coefficients for CH₄ mixing ratios filtered with the 5th percentile exceeded $R = 0.7$ in 67 %
514 of all cases. This means that in the majority of cases more than ≈ 50 % (derived from
515 $R^2 \geq 0.7^2$) of the variance in the CH₄ mixing ratios at the mountain top could be explained
516 by the concurrent tall-tower measurements and vice versa.

517 The WDS filter led to similarly high correlations between the mountain-top and the
518 tall-tower station CH₄ mixing ratios ($R \geq 0.7$ in 67 % of the cases) as the 5th percentile
519 (Fig. 10 b). Weaker correlations ($R < 0.7$) could often be attributed to periods with
520 weakly varying atmospheric CH₄ mixing ratios at both stations. Usually periods with
521 very high correlations were also characterised by offsets between the stations being close
522 to zero. The good overall agreement between both filtered time series indicates that CH₄
523 mixing ratios at the mountain-top station follow the same seasonal trend as CH₄ mixing
524 ratios at the tall-tower station, but it was not possible to completely eliminate the offset at
525 the mountain-top station imposed by local sources and/or mountain flows, even after
526 screening 2-h averages with a WDS filter or filtering the CH₄ data using the 5th percentile
527 approach.

528 Wintertime CO₂ mixing ratios at the two stations agreed rather well after applying the 5th
529 percentile or the WDS filter, as is indicated by the high correlation coefficients (typically
530 $R > 0.7$). From January to April, there was an excellent agreement of CO₂ mixing ratios
531 between the tall-tower and the mountain-top station (Fig. 10 c, d). During the growing
532 season, the relationship was weakest (usually with $0.3 < R < 0.6$, Fig. 10 d). During
533 winter the relationship between 5th percentiles of the CO₂ mixing ratios measured at both

534 stations was as good as the relationship between the WDS filtered 2-h average mixing
535 ratios, whereas in summer the WDS filtering usually did not retain any values.

536



537

538 **Fig. 10** Comparison of the CH₄ (a) and CO₂ (c) mixing ratios at the mountain-top (green crosses) and the
539 tall-tower (blue asterisk) station after the application of the WDS filter. Pearson's correlation coefficients (*R*)
540 between 5th percentiles of mountain-top and tall-tower measurements (black solid line) and 2-h averages of
541 mountain-top and tall-tower measurements retained after WDS filtering (blue dots) calculated within a time
542 window of ± 7 days around every measurement point are shown for CH₄ (b) and CO₂ (d), respectively. The
543 Pearson correlation coefficients are all significantly different from zero ($p < 0.01$). Management activities
544 affecting the surrounding meadows are highlighted in grey in panels b and d. The management activities at
545 the mountain-top station were: application of organic fertilizer or grazing of cattle on the pastures
546 surrounding the measurements (b) and harvests (d).

547

548 **4 Discussion**

549 Precise measurements of CO₂ and CH₄ mixing ratios have been carried out at several
550 stations in Europe and North America already (e.g. Thompson et al. 2009, Göckede et al.
551 2010, Sun et al. 2010, Winderlich et al. 2010, Vermeulen et al. 2011, Lavaux et al. 2012,
552 Miles et al. 2012, Andrews et al. 2014). To the best of our knowledge, however, this
553 study is the first to discuss the measurements obtained at two stations in a paired-site
554 approach, comparing a tall-tower with a neighbouring (≈ 28 km separation) mountain-top
555 station at a similar elevation above sea level. Since local effects are usually not
556 represented well in atmospheric transport models it is essential to extract the regional
557 signal from mountain-top measurements. Thus, the key challenge for inverse modelling
558 in the comparison between tall-tower and mountain-top greenhouse gas mixing ratios is
559 the filtering of the dataset to remove local influences that only affect one but not the other
560 site. On the other hand, meteorologists trying to qualitatively and quantitatively
561 understand such local effects in complex terrain are typically interested in periods or in
562 meteorological conditions when differences between two sites are largest. In order to
563 address both aspects, we first discuss the extraction of the regional signal using the
564 filtering approach suggested in Section 3.3, and then address the meteorological
565 differences between the sites under conditions when local effects play an important role.

566

567 4.1 Extracting the Regional Signal

568 Two approaches were used to extract the regional signal from the time series of mixing
569 ratio measurements: (i) using a statistical percentile filter, and (ii) using a conditional
570 filter based on wind direction, wind speed, time of day, and season (WDS filter). An
571 alternative filtering approach, which was already used to screen data from well mixed air
572 masses, is time-of-day filtering (Göckede et al. 2010, Peters et al. 2010), which
573 exclusively uses well mixed afternoon conditions and nighttime measurements in the
574 residual layer well above the nocturnal boundary layer. More sophisticated statistical
575 approaches which have been used to filter mountain-top measurements are (i) short-term
576 variance filters and (ii) weighted median smoothers (Brooks et al., 2012). In our case, the
577 5th percentile approach proved to be a suitable choice to filter outliers related to local CH₄
578 sources. However, this approach makes the implicit assumption that within a 2-hour time

579 period the lower-than-average mixing ratios are most likely closest to the regional signal,
580 while higher-than-average mixing ratios are more likely to be outliers caused by local
581 sources. Using a 5th percentile filter led to rather robust comparisons between tall-tower
582 and mountain-top sites for CH₄ mixing ratios (Fig. 5). Its shortcomings, however, are its
583 relatively strong sensitivity to local sinks, its statistical and non-causal basis, and the fact
584 that there is no clear threshold that would establish the 5th percentile as the best choice
585 everywhere. Contrastingly, the second approach – the WDS filter – uses conditional
586 criteria that are easier to understand in the context of impacts of the local topography and
587 local surface fluxes. For CH₄, both filtering approaches seem to be useful, whereas in the
588 case of CO₂ none of the filters worked well during the growing season (Fig. 10),
589 indicating that the regional signal is too strongly altered by local influences to be
590 extracted with a WDS filter.

591

592 4.2 The Importance of Local Effects

593 *4.2.1 Terrain Driven Effects on Greenhouse Gas Mixing Ratios*

594 The accumulation of CH₄ and other trace gases in the nocturnal boundary layer at valley-
595 bottom sites during the night is a general pattern found in diurnal cycles of mixing ratio
596 measurements and has been used before to determine regional-scale CH₄ fluxes (e.g.,
597 Stieger et al. 2015). At the mountain top, cold air drainage flows prevent the build-up of a
598 deep, stable nocturnal boundary layer and thus the boundary layer extends 200–300 m
599 above the valley bottom, even in broad valleys such as the Swiss Plateau (cf. Eugster and
600 Siegrist, 2000; see also Stieger et al., 2015). Hence, although our mountain-top
601 measurements were only performed at 4-m height above ground surface, the diurnal
602 courses of CH₄ and CO₂ mixing ratios rather followed that of the tall-tower site (Figs. 5
603 and 6) than that of a valley-bottom site (Fig. 7). This indicates that the mountain-top
604 station – similarly to the tall-tower station – remains above the stable nocturnal boundary
605 layer of the surrounding valleys, where agriculturally driven methane emissions lead to a
606 substantial increase of methane during the night (Fig. 7). Consequently, both sites were
607 decoupled from the conditions at the valley floor during the night. This agrees well with
608 mobile measurements that were performed along the slopes of the Reuss valley and up to

609 the Frübüel mountain-top site (Bamberger et al. 2014) which clearly showed that the
610 nocturnal boundary-layer depth is only between 100 m and 200 m deep during clear
611 summer days, while the Frübüel mountain-top site is located almost 500 m above the
612 valley floor. Considering the low sampling height above ground, however, the mountain-
613 top site is located in the local surface layer.

614 Greenhouse gas mixing ratios at the tall-tower site showed a well-known increase of
615 mixing ratios in the late morning with the onset of daytime convective mixing (Davis et
616 al. 2003) which was virtually absent at the mountain-top site. The absence of the late
617 morning peak at the mountain top can be explained by its location on top of a mountain
618 ridge on the easterly side of the Reuss valley (Bamberger et al. 2014). The onset of
619 daytime convective mixing and vertical advection of nighttime pollution along the
620 warmer hillslopes is a pattern that is common to mountainous regions (Gohm et al. 2009,
621 Schnitzhofer et al. 2014). Considering the diurnal cycle of greenhouse gas mixing ratios,
622 the tall-tower station (nominally 212 m a.g.l.) also profits from its position on a local hill
623 that removes the tower top from the valley bottom (where CH₄ and CO₂ accumulates
624 during the night) by an additional ≈ 300 m: the tower base is at 797 m a.s.l., whereas the
625 flat areas surrounding the locality are at 463 m (Lake Baldegg) to 504 m a.s.l. (Lake
626 Sempach). Without this additional topographic height, more pronounced diurnal cycles in
627 CH₄ and CO₂ mixing ratios would be expected at the tall-tower site: Winderlich et al.
628 (2014) found that at the highest level (301 m above ground) of the Zotino tower, Siberia,
629 occasional nocturnal increases in CH₄ mixing ratios occur.

630 All these considerations suggest that for both mountain-top and tall-tower sites it is not
631 the height of a measurement from the local ground that matters, but the larger-scale
632 topographic context, and thus the altitude above the topographic reference elevation,
633 which is typically the bottom of a larger and broader valley where nocturnal
634 accumulation of greenhouse gases takes place.

635 *4.2.2 Effects of Local Flow*

636 Flows at the tall-tower and the mountain-top stations showed both a strong channelling
637 along the two main wind directions (north-east and south-west) and minor contributions
638 from other directions, mostly in combination with low wind speeds. On the other hand,

639 there was a directional shift between the main wind directions at the mountain-top site
640 (which was turned anti-clockwise) as compared with the tall-tower site. However, such a
641 rotation of the flow between the surface and the mid-boundary layer is well known as the
642 Ekman spiral (e.g. Holton 2004). Due to the well-known natural increase of wind speed
643 with height above surface, wind speeds within the surface layer measured at 2 m a.g.l. at
644 the mountain-top site were considerably lower than those at the tall-tower site (at 212 m
645 a.g.l.) In addition, wind speeds at the tall-tower site were maximal during the night,
646 whereas they were lowest during the night at the mountain-top site, at least during
647 summer (Oney et al. 2015).

648 At the mountain top, flow from south-easterly direction was occasionally stronger (up to
649 10 m s^{-1}) than flow from other directions. These high wind speeds usually occurred in
650 combination with high temperatures and persistent wind directions (data not shown) and
651 indicated the influence of foehn winds (warm and dry southerly winds across the Alps).
652 The influence of foehn winds at the mountain-top station is plausible as it is located near
653 the main axis of the Reuss Valley, one of the main foehn valleys in Switzerland (Seibert
654 1990), whereas the tall-tower station is much more sheltered against the southerly flow
655 by Mount Rigi and other mountain ridges and is further away from the Alps. At the
656 mountain-top station, wind directions outside the usual easterly or westerly directions
657 were more frequent than at the tall-tower station but generally weak (mostly $< 2 \text{ m s}^{-1}$),
658 suggesting thermally induced south-westerly flows along the slopes of the Zugerberg
659 mountain ridge and larger scale northerly plain-to-mountain flows (Lugauer and Winkler
660 2005), a phenomenon commonly referred to as Alpine pumping. While easterly (down-
661 slope) winds were seen predominantly during night-time, westerly (up-slope) winds were
662 more frequent during the day, agreeing with the typical picture of diurnal mountain winds
663 (Whiteman 2000) in the surface layer. These findings agree with the findings by Oney et
664 al. (2015) that the local environment influences prevailing wind directions more at the
665 mountain-top station than at the tall-tower station.

666 *4.2.3 The Effect of Local Sources and Sinks*

667 During time periods when the ETH research station and its farmstead buildings were in
668 upwind direction of the mountain-top site the offset between both stations was 0.065 ppm

669 on average, whereas it was much lower (< 0.01 ppm) for all other wind directions (see
670 Fig. S1). The small directional deviation between the sector with highest CH_4 mixing
671 ratios and the farmstead buildings (seen in Fig. S1) relates to the influence of local terrain.
672 The close farmstead obviously acted as a source of CH_4 which considerably influenced
673 the mixing ratios when the flow came from that direction (Figs. S1 and 8). With higher
674 wind speeds and south-easterly flow, CH_4 and CO_2 mixing ratios at the mountain-top site
675 were typically lower than those at the tall-tower site (Figs. 8 and 9). This indicates that
676 the mountain-top station observes relatively clean free-tropospheric air masses
677 descending in the lee of the Alps during foehn as opposed to the tall-tower station
678 measuring planetary boundary-layer air. Since foehn wind tends to follow the terrain
679 similarly to hydraulic flows, our assumption that the same elevation above sea level can
680 be compared fails under such special conditions where altitude above the topographic
681 reference surface becomes more relevant than the absolute elevation. Additionally, during
682 nighttime and in winter, when vertical mixing is generally weaker, mixing ratios at wind
683 speeds $< 2 \text{ m s}^{-1}$ were often larger at the mountain-top as compared to the tall-tower
684 station, especially with westerly up-slope winds and northerly winds which most likely
685 originated from lower parts of the valley. Thus, with the WDS filter all conditions with (a)
686 flow from the farmstead, (b) wind directions and speeds associated with foehn, and (c)
687 during times with reduced vertical mixing, wind speeds $< 2 \text{ m s}^{-1}$ and northerly or
688 westerly wind directions were removed from the time series. When wind speeds are low
689 ($0\text{--}1 \text{ m s}^{-1}$), wind directions are often not properly defined or very variable. This could
690 explain why CH_4 mixing ratios at wind speeds $< 1 \text{ m s}^{-1}$ are higher at the mountain top
691 even if wind direction does not include the local farmstead. Moreover, a strong northerly
692 or westerly component in the flow means that the air mass passed the Zugerberg
693 mountain ridge where other farmsteads are located. During periods of low solar radiation,
694 weak turbulence, and low wind speeds, the influences from more remote farmsteads
695 could be increased, which could explain the higher CH_4 mixing ratios observed at the
696 mountain-top station.

697 Periods when CO_2 mixing ratios were higher at the mountain-top station than those at the
698 tall-tower site did not show such pronounced wind-direction dependence as did CH_4 . In
699 spring and summer/autumn, the difference between CO_2 mixing ratios at the mountain-

700 top and the tall-tower station was typically highest with lower wind speeds $< 2 \text{ m s}^{-1}$ and
701 reduced vertical mixing (evening, night). As the vegetation surrounding the mountain-top
702 site is covered by grassland and forests, it acted both as a source (respiration) and a
703 strong sink (photosynthesis) of CO_2 depending on time of day or season and management
704 (Zeeman et al. 2010). This generally limited the capability to distil the anthropogenic
705 CO_2 signal out of the CO_2 mixing ratio time series. Here an increased measurement
706 height above the local surface may help to reduce the strong influence from the
707 vegetation (e.g. Bakwin et al. 1995, 1998). Alternatively, there are methods to estimate
708 regional fluxes based on vertical gradients of mixing ratios between the atmospheric
709 boundary layer and the free troposphere at tower sites (e.g. Bakwin et al. 2004, Crevoisier
710 et al. 2006). In a well-mixed atmospheric boundary layer, flux–gradient relationships
711 (Monin and Obukhov 1954, Moeng and Wyngaard 1984, 1989) have been used to adjust
712 for biases introduced by low measurement heights (Bakwin et al. 2004). In winter, CO_2
713 mixing ratios at the mountain-top were often higher than at the tall-tower site with flows
714 from the farmstead and flows from northerly or westerly directions, similarly to CH_4
715 mixing ratios. These high biases were most frequent during periods of reduced vertical
716 mixing and thus were removed from our analysis.

717 In summary, our comparison between a tall-tower site and a mountain-top site provided
718 evidence that the regional signal of CH_4 mixing ratios can be extracted from the time
719 series even under the presence of geographically constrained local sources. However, a
720 bias of approximately 0.01 ppm in the CH_4 mixing ratio has to be taken into account. It
721 remains much more challenging to do the same with CO_2 mixing ratios due to the
722 dominance of the local biogenic signal present in the measurements when plant
723 photosynthesis and ecosystem respiration are most active.

724

725 **5 Conclusions**

726 Atmospheric CO_2 and CH_4 observations at a mountain-top site have been compared to
727 measurements at a neighbouring tall-tower site (within 28.4 km horizontal distance) at a
728 similar elevation above sea level. Although the airflow was significantly perturbed
729 locally at the mountain-top site, CH_4 mixing ratios were quite similar to those at the tall-

730 tower site, except for peak values. Average mountain-top and tall-tower CH₄ mixing
731 ratios showed good agreement, with an average CH₄ bias between the two stations that
732 was around 0.01 ppm, after applying a filter to select for favorable wind direction, wind
733 speeds, time of day and season at the mountain-top site, and 5th percentiles of the 2-h
734 frequency distribution of CH₄ mixing ratios at both sites. Peak mixing ratios of unfiltered
735 data, however, were clearly influenced more by local CH₄ sources at the mountain top
736 site than at the tall-tower site. Consequently, it was possible to minimize, but not to
737 completely remove, the influence of local agricultural CH₄ sources by choosing an
738 appropriate filter. Hence, we conclude that, at least in absence of local sources, a
739 mountain-top station can provide greenhouse gas observations with similar regional
740 representativeness as a tall-tower station. Geographically well-defined CH₄ sources may
741 be acceptable to a certain degree as their influence can be removed with appropriate
742 filtering. However, it is generally preferable to choose the location of mountain-top sites
743 in a way that local sources are virtually absent.

744 For CO₂, however, the usefulness of mountain-top mixing ratio measurements may be
745 more limited for inverse modelling, especially during summer, when vegetation and soils
746 cause a more pronounced diurnal cycle at the mountain top than on the top of a tall tower.
747 Since these diurnal and seasonal signals are strong, a filter based on local wind conditions,
748 time of day, and season is not able to remove all local influences of the vegetation at the
749 mountain-top site. At sites where the larger-scale (far-field) signal to local-scale noise
750 ratio remains an issue, an increased measurement height should be considered, leading to
751 a dilution and damping of the local-scale noise. Being aware of such limitations, we
752 conclude that a carefully selected mountain-top site still can be considered a suitable
753 alternative for a tall-tower station. Mountain sites have a similar potential for continuous
754 long-term monitoring, and could complement tall-tower stations, especially in
755 mountainous terrain.

756

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Supplementary Material: Observations of atmospheric methane and carbon dioxide mixing ratios: Tall-tower or mountain-top stations?

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S1 Management Activities at the Mountain-Top Site

Table S1 Management activities at the grasslands surrounding the FRU mountain-top station

Date in 2013	Management activity	Details
17–18 April	organic fertilizer	solid
2–8 May	grazing	suckler cows (25), calves (24)
6–14 June	grazing	cows (34)
1–2 July	cutting	0.68 ha
11–12 July	cutting	0.71 ha
18 July	organic fertilizer	solid
23 July	cutting	3.28 ha
24 July	organic fertilizer	liquid
2–4 September	cutting	0.68 ha
23 September–2 October	grazing	cows (34)
7–13 November	grazing	sheep (15)

S2 Origin of Methane Emissions

10 To determine whether the local methane sources at the mountain-top station are diffuse or attributable to a specific source we analyzed the dependence of the between-site differences ($\Delta C_m = C_{m,FRU} - C_{m,BER}$) on the wind direction at the mountain top (FRU). This was done both for two-hourly average CH₄ mixing ratios and for the 5th percentiles of the two-hourly frequency distribution (Fig. S1). For average CH₄ mixing ratios these differences remained generally below 0.02 ppm
 15 for any wind direction, except for southerly winds from the direction of the ETH research station when average mixing ratios at the FRU were considerably higher (up to 0.065 ppm).

This pattern was not restricted to dates when cattle were grazing on the pastures surrounding the FRU measurement station, thus indicating the general influence of CH₄ emissions from the nearby farmstead in the south. Evaluating the between-site differences for the 5th percentiles instead of
 20 the average mixing ratios, however, showed no comparable dependence on wind direction (Fig. S1). Thus the percentile approach, which assumes that the lower edge of the two-hourly frequency distribution is mainly unaffected by local greenhouse gas fluxes, was successful in eliminating the most prominent influences on CH₄ mixing ratios imposed by the local farmstead.

S3 Analyzing Mixing Ratio Differences

25 In order to find the most likely predictor variables that influence the mixing ratio difference between the FRU (mountain-top) and BER (tall-tower) sites, we carried out a principal component analysis

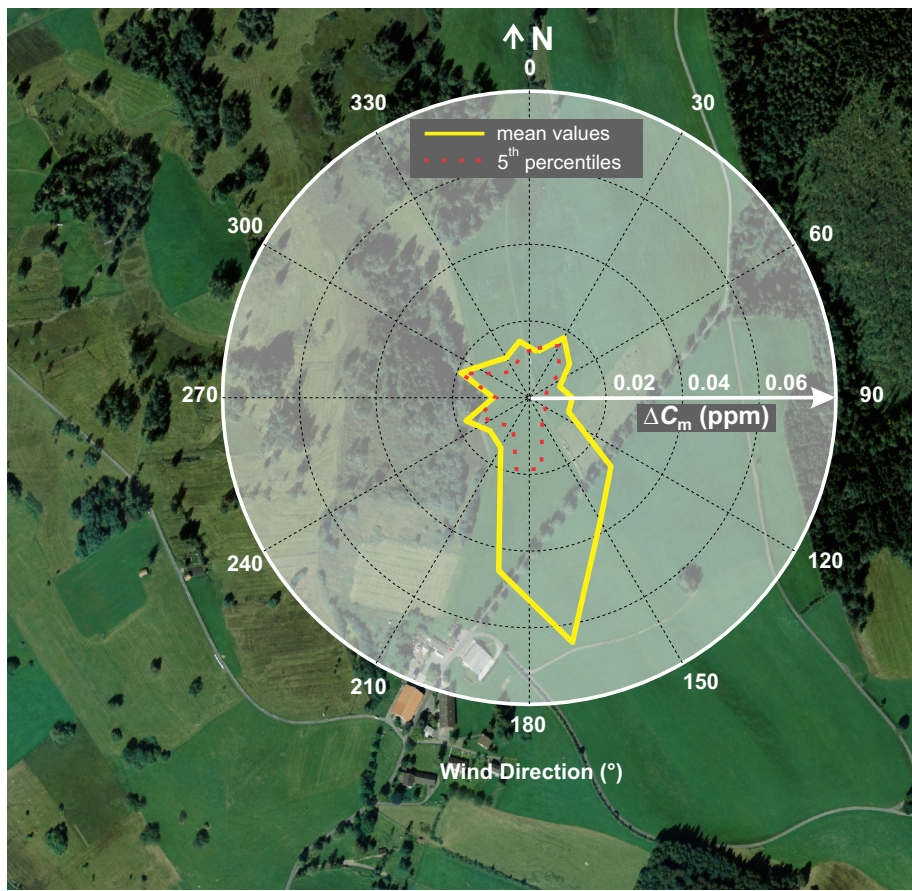


Fig. S1 Binned averages of the CH_4 differences between the FRU mountain-top station and the BER tall-tower station (ΔC_m) as calculated from the annual time series of 2-h averages (yellow line) and the 5th percentiles of the 2-h frequency distribution (red dotted line) for 20° bins of the wind direction at the mountain top, overlaid on an orthoimage © 2014 swisstopo (JD100042) of the mountain-top site

(not shown) to detect collinearities among the meteorological variables that have the potential to predict such differences, and then performed an analysis of variance to determine, which groups of environmental conditions can be used to classify mixing ratio differences between FRU and BER.

30 All statistical analyses were done with R (R Core Team, 2016).

The response variables of interest were ΔC_m and ΔC_c , the absolute differences in CH_4 and CO_2 mixing ratio measurements, respectively, measured at the two sites. Positive values of ΔC_m and ΔC_c indicate that the respective gas mixing ratio measured at the FRU mountain-top site exceeded the concurrent measurements at the top of the BER tall-tower site.

35 S3.1 Analysis of Variance

S3.1.1 Material and Method

We carried out an analysis of variance to quantify the differences between the FRU and BER mixing ratio measurements. The goal was not primarily to determine the offset between the sites, but how this potential offset varies with the variables used in the principal component analysis. We used a three-step procedure that takes into account that different variables are most relevant for ΔC_m than for ΔC_c :

1. A five-factorial analysis of variance was computed with the variables wind direction (ϑ), horizontal wind speed (U), atmospheric stability (z/L ; all measured at FRU), H_r (hour of day),

and S_n (season, starting with 12 levels corresponding to calendar months). To simplify calculations, all variables were binned to classes to reduce the number of levels. In the first step no interactions between the different variables were considered. The H_r variable was binned at 2-h resolution, and atmospheric stability was aggregated to the three classes “unstable” ($z/L < -0.0625$), “neutral” ($-0.0625 \leq z/L < 0.0625$), and “stable” ($z/L \geq 0.0625$), since both z/L and L/z only showed a weak effect in both principal component analyses for ΔC_m and ΔC_c . Wind speed was binned into classes at intervals of 1 m s^{-1} , and ϑ (as measured at the FRU site) was binned in 10° segments. Tukey’s honest significant difference test was calculated on all groupwise comparisons of mixing ratio differences.

2. After this, the classes without significant in-between differences were combined to reduce the number of classes in each variable. U was aggregated to five classes ($0\text{--}1 \text{ m s}^{-1}$, $1\text{--}2 \text{ m s}^{-1}$, $2\text{--}5 \text{ m s}^{-1}$, $5\text{--}6 \text{ m s}^{-1}$, $> 6 \text{ m s}^{-1}$ for ΔC_m , and $< 1 \text{ m s}^{-1}$, $1\text{--}2 \text{ m s}^{-1}$, $2\text{--}4 \text{ m s}^{-1}$, $4\text{--}6 \text{ m s}^{-1}$, $> 6 \text{ m s}^{-1}$ for ΔC_c). ϑ was aggregated to three classes ($70\text{--}120^\circ$, $120\text{--}200^\circ$, $200\text{--}70^\circ$). To use a more detailed resolution of H_r and S_n than in Fig. 8, an aggregation was done as shown in Fig. S2 with four classes each for H_r and S_n .
3. With this reduced set of classes a full analysis of variance model including interactions was run. Using stepwise exclusion, all nonsignificant variables and interactions (adjusted $p \geq 0.05$) were subsequently eliminated from the model. Within a few iterations we were left with a simplified analysis of variance model that only contained variables and interactions that showed a nonzero influence on measured ΔC_m or ΔC_c , respectively, at adjusted $p < 0.05$ (see Tables S2 & S3).

S3.1.2 Results

On average the mixing ratio offset between FRU and BER was 0.019 ± 0.019 ppm for ΔC_m , and 2.9 ± 3.7 ppm for ΔC_c (mean difference \pm standard deviation of all group comparison of the groups that were used in the final analysis of variance model).

Differences in CH₄ mixing ratios.

In the first step (no interactions considered) an aggregation of predictor classes was possible, namely H_r could be reduced to two classes: 0800–1600 UTC (day), and 2000–0400 UTC (night). Transition times were not included in the analysis. The 12 months of S_n could be aggregated to four seasons, January, February–July, August, and September–December. January and August differed in all comparisons with other months, hence this site-specific classification was chosen (Fig. S2). Wind speed classes could be reduced to five classes, and ϑ was aggregated to three classes (see above). The two stability groups “unstable” and “neutral” did not differ significantly and thus were combined. These classes were used in step 3 to obtain the final analysis of variance Table S2.

The output from the final analysis of variance model (Table S2) was then used to compute group averages, standard deviations, and standard error of the mean of ΔC_m to quantify the possible local effects on ΔC_m . These averages were then ranked and plotted in Fig. S3 in combination with the predictor classes involved.

Figure S3 shows quite clearly that wind directions from $120\text{--}200^\circ$ were mostly responsible for the largest positive deviations, i.e., local emissions at FRU increase CH₄ mixing ratios above the value expected at BER. Contrastingly, the most negative deviations are associated with the highest wind speeds and with the early season (February to July).

Wind directions from $70\text{--}120^\circ$ (Bise winds; Wanner and Furger, 1990) never led to significantly increased ΔC_m compared to other factor combinations, and hence those wind directions can be considered representative for BER conditions, irrespective of H_r , S_n , U , and z/L . Whereas stability, U , and H_r had a similar influence on the variability of ΔC_m (F ratios in the range 40–50, Table S2), only U showed a distinct influence on average ΔC_m . While weak winds led to positive ΔC_m , strong winds were usually associated with negative ΔC_m . Thus, U – which is also related to stability – had a much stronger influence than stability on local anomalies in CH₄ mixing ratio measurements at FRU.

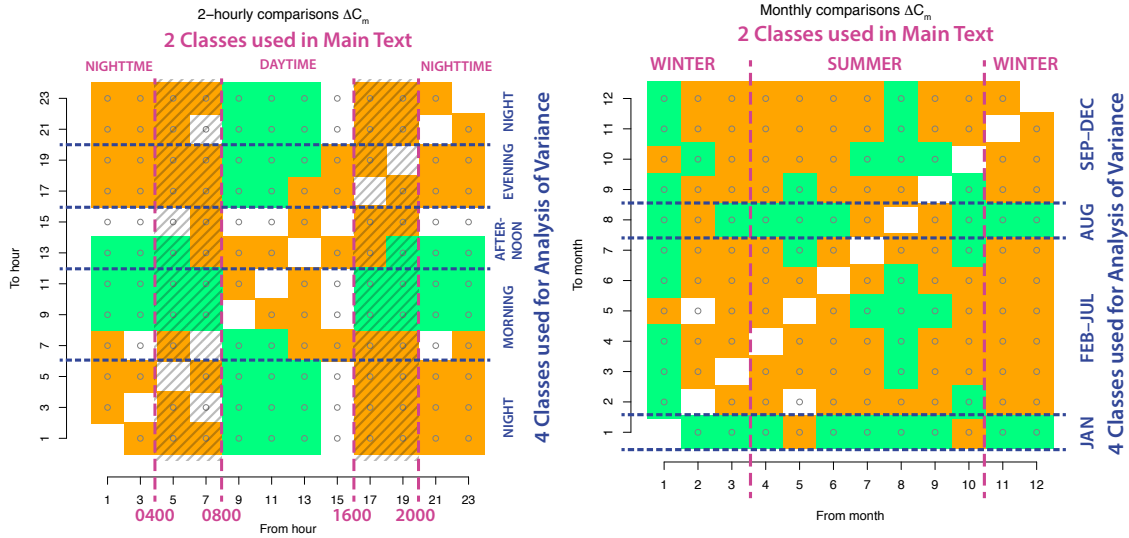


Fig. S2 Aggregation of hour of day (H_r , left) and season (S_n , right) for the analysis of variance of CH_4 mixing ratio differences. Both panels show a graphical matrix of pairwise comparisons of CH_4 mixing ratio differences. Orange cells denote that Tukey's honest significant difference test yielded an adjusted $p > 0.05$, whereas green cells show highly significant differences (adjusted $p < 0.01$). White cells are either significant in the range $0.01 \leq \text{adjusted } p \leq 0.05$, or are the diagonal of the matrix which is ignored. The vertical dashed lines show the coarse separation of the classes used in Fig. 8, whereas the horizontal dashed lines show the separations between classes used in this additional analysis of variance (Fig. S3 and Table S2). The hashed areas are the transition times that are excluded in the analysis for Fig. 8, but are included in the analysis of variance Table S2

Mean ΔC_m was 0.019 ± 0.019 ppm, shown in Fig. S3 with the two dashed vertical lines. The most positive ΔC_m ranged up to 0.060 ppm (January, neutral z/L , calm winds $< 1 \text{ m s}^{-1}$, and ϑ 120–200°), and the most negative ΔC_m was -0.045 ppm (afternoons with relatively strong winds with U in the range 7–8 m s^{-1}).

Differences in CO_2 mixing ratios.

For ΔC_c the aggregation of the variables led to slightly different classes than for ΔC_m . For simplicity we used the same H_r classes as for CH_4 (night 2000–0600 UTC, morning 0600–1200 UTC,

Table S2 Analysis of variance table for the response variable ΔC_m after stepwise elimination of nonsignificant predictors and predictor combinations, sorted by F value

Factor Interaction	Df	Sum Sq	Mean Sq	F value	$\text{Pr}(> F)$	
ϑ	2	0.6538	0.32691	249.6591	$< 2.2 \cdot 10^{-16}$	***
z/L	1	0.0648	0.06476	49.4552	$2.398 \cdot 10^{-12}$	***
U	7	0.3946	0.05638	43.0564	$< 2.2 \cdot 10^{-16}$	***
H_r	3	0.1575	0.05249	40.0872	$< 2.2 \cdot 10^{-16}$	***
S_n	3	0.1117	0.03724	28.4392	$< 2.2 \cdot 10^{-16}$	***
$\vartheta : U$	6	0.0728	0.01213	9.2623	$4.227 \cdot 10^{10}$	***
$\vartheta : z/L$	2	0.0211	0.01054	8.0496	0.0003247	***
$S_n : H_r$	9	0.0477	0.00530	4.0448	$3.561 \cdot 10^{-5}$	***
$\vartheta : H_r$	6	0.0249	0.00415	3.1680	0.0042138	**
$\vartheta : S_n : U$	23	0.0703	0.00306	2.3333	0.0003198	***
$\vartheta : H_r : z/L$	9	0.0238	0.00265	2.0236	0.0330805	*
$\vartheta : S_n : H_r$	23	0.0504	0.00219	1.6732	0.0230944	*
Residuals	3783	4.9535	0.00131			

Asterisks in the last column indicate $p < 0.001$ (***), $p < 0.01$ (**), and $p < 0.05$ (*), and the colons in the first column indicate interactions among variables. The following classes were built for each variable: wind direction ϑ 70–120°, 120–200°, 200–70°; wind speed $U < 1 \text{ m s}^{-1}$, 1–2 m s^{-1} , 2–5 m s^{-1} , 5–6 m s^{-1} , $> 6 \text{ m s}^{-1}$; hour of day H_r 0600–1200 (morning), 1200–1600 (afternoon), 1600–2000 (evening), 2000–0600 (night); season S_n January, February–July, August, September–December; stability z/L stable, unstable and neutral.

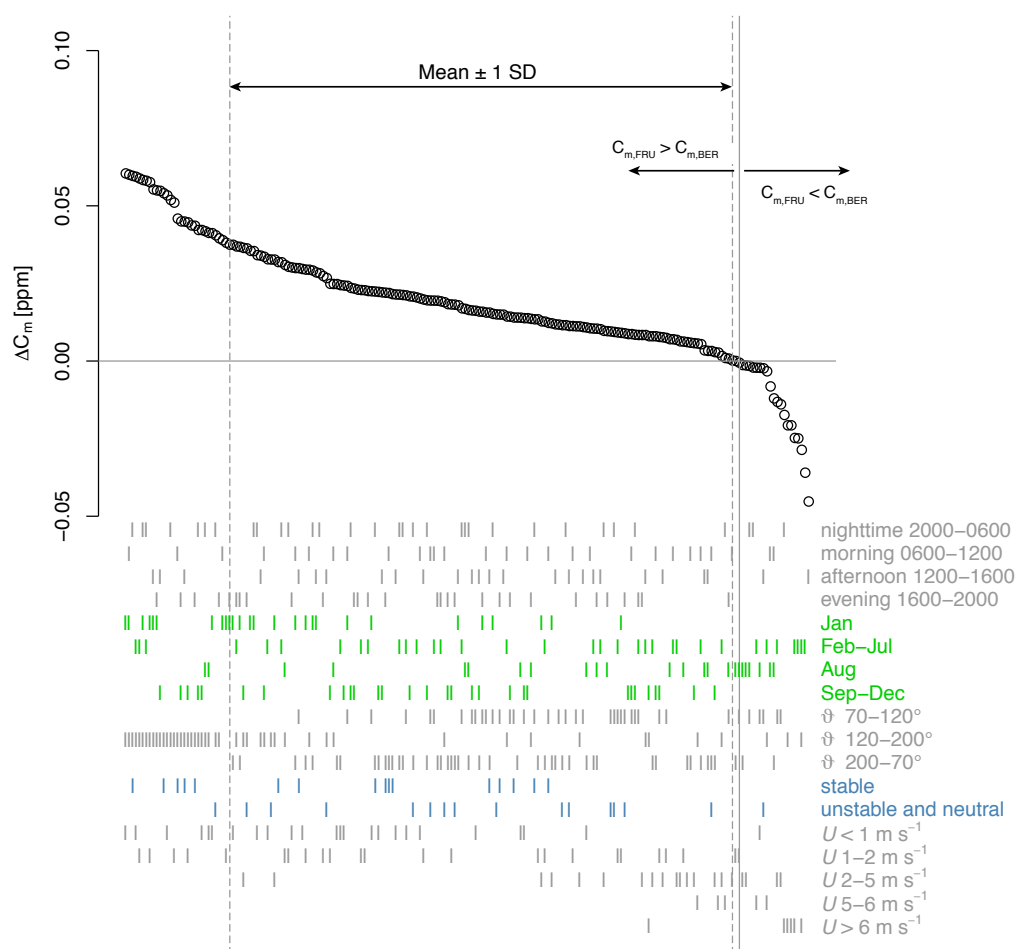


Fig. S3 Group averages of differences in CH₄ mixing ratios measured at the FRU site in comparison with the BER site, sorted in descending order. The mean difference \pm one standard deviation (i.e., 0.019 ± 0.019 ppm) is shown with two vertical dashed lines. The zero difference is shown with a horizontal and a vertical solid thin line. The vertical symbols below the graph indicate which class combination was used for the respective group average. As a reading example: the most positive values of ΔC_m are found in group comparisons that involve the wind direction sector 120–200°, and the most negative deviations are found at the highest wind speeds

100 afternoon 1200–1600 UTC, evening 1600–2000 UTC), and the transition times (0400–0800 UTC and 1600–2000 UTC) were excluded from further analysis (Fig. S4) in the same way as was done with CH₄. Seasonality was represented by four levels: November–April, May–June, July–August, and September–October. For ϑ we used the same three sectors as for CH₄: ϑ 70–120°, 120–200°, 200–70°. For U five levels were used: $< 1 \text{ m s}^{-1}$, $1\text{--}2 \text{ m s}^{-1}$, $2\text{--}5 \text{ m s}^{-1}$, $5\text{--}6 \text{ m s}^{-1}$, and $> 6 \text{ m s}^{-1}$.
 105 Stability had three levels: “unstable”, “neutral”, and “stable”. The final analysis of variance model is shown in Table S3.

Similarly to ΔC_m we sorted the differences and plotted the values along with the classes involved (Fig. S5). Beginning at the bottom of Fig. S5 it is quite clearly seen that large positive differences are never associated with wind speeds above 2 m s^{-1} , and the most negative deviations are associated with high wind speeds (and thus enhanced turbulent mixing). At the same time, the most negative values of ΔC_c tend to be related to the southeastern wind sector. This wind sector is mostly associated with foehn conditions (Desai et al., 2016), that bring fresh and warm air from the Alps directly to FRU, whereas the BER site actually should already see parts of the anthropogenic emissions in the Alpine foreland. Hence such differences—although weak and only represented by few datapoints—are realistic and most likely true differences, not sampling artefacts. Very positive deviations are primarily associated with nighttime or early morning (dusk), and mostly during
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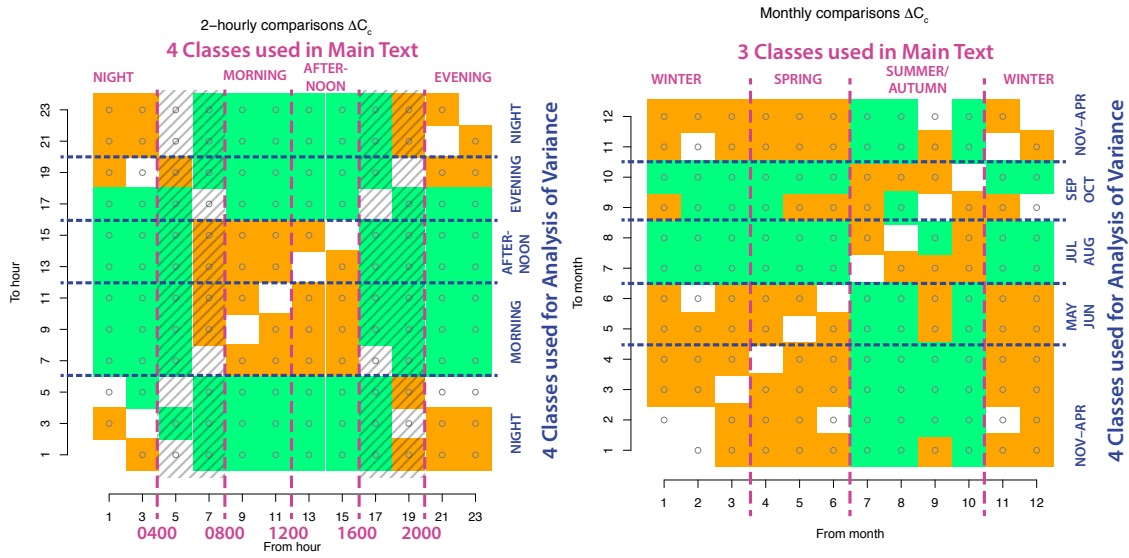


Fig. S4 Aggregation of hour of day (H_r , left) and season (S_n , right) for the analysis of variance of ΔC_c . Both panels show a graphical matrix of pairwise comparisons of ΔC_c mixing ratio differences. Orange cells denote that Tukey's honest significant difference test yielded an adjusted $p > 0.05$, whereas green cells show highly significant differences ($p < 0.01$). The vertical dashed lines show the coarse separation of the classes used in Fig. 9, whereas the horizontal dashed lines show the separations between classes used in this additional analysis of variance (Fig. S5 and Table S3). The hashed areas are the transition times that are excluded in the analysis of variance Table S3. White cells are either the diagonal of the matrix or differences in the range $0.01 < p \leq 0.05$

Table S3 Analysis of variance table for the response variable ΔC_c after stepwise elimination of nonsignificant predictors and predictor combinations, sorted by F value

Factor Interaction	Df	Sum Sq	Mean Sq	F value	$\Pr(> F)$	
H_r	3	20553	6851.2	363.5365	$< 2.2 \cdot 10^{-16}$	***
ϑ	2	8716	4357.8	231.2329	$< 2.2 \cdot 10^{-16}$	***
U	4	7569	1892.3	100.4102	$< 2.2 \cdot 10^{-16}$	***
S_n	3	5265	1755.2	93.1325	$< 2.2 \cdot 10^{-16}$	***
z/L	2	2767	1383.3	73.4019	$< 2.2 \cdot 10^{-16}$	***
$S_n : H_r$	9	7604	844.9	44.8299	$< 2.2 \cdot 10^{-16}$	***
$\vartheta : S_n$	6	4141	690.1	36.6186	$< 2.2 \cdot 10^{-16}$	***
$H_r : z/L$	6	699	116.5	6.1814	$1.823 \cdot 10^{-6}$	***
$H_r : U$	12	893	74.4	3.9496	$4.379 \cdot 10^{-6}$	***
$S_n : U$	11	778	70.8	3.7547	$2.296 \cdot 10^{-5}$	***
$\vartheta : H_r$	6	376	62.6	3.3243	0.002882	**
$\vartheta : S_n : H_r$	16	710	44.4	2.3547	0.001751	**
$\vartheta : S_n : U$	19	768	40.4	2.1435	0.002731	**
$S_n : H_r : z/L$	16	615	38.4	2.0390	0.008491	**
$\vartheta : H_r : z/L$	12	436	36.3	1.9285	0.026884	*
$\vartheta : S_n : z/L$	22	794	36.1	1.9150	0.006244	**
$\vartheta : S_n : z/L : U$	51	1564	30.7	1.6275	0.003339	**
Residuals	3689	69522	18.8			

Asterisks in the last column indicate $p < 0.001$ (***), $p < 0.01$ (**), and $p < 0.05$ (*), and the colons in the first column indicate interactions among variables. The following classes were built for each variable: wind direction ϑ 70–120°, 120–200°, and 200–70°; wind speed $U < 1 \text{ m s}^{-1}$, 1–2 m s^{-1} , 2–4 m s^{-1} , 4–6 m s^{-1} , > 6 m s^{-1} ; hour of day H_r 0600–1200 (morning), 1200–1600 (afternoon), 1600–2000 (evening), 2000–0600 (night); season S_n November–March, April–June, July–August, September–October; stability z/L stable, neutral, unstable.

the warm season from July until October. This coincides with the period when the vegetation is most active during the day. At the same time soils are warm enough to enhance the activity of soil microbes that decompose soil organic matter and increase the local CO_2 mixing ratio, namely at night. The largest positive ΔC_c observed was 14.3 ppm (July–August, neutral z/L , calm winds

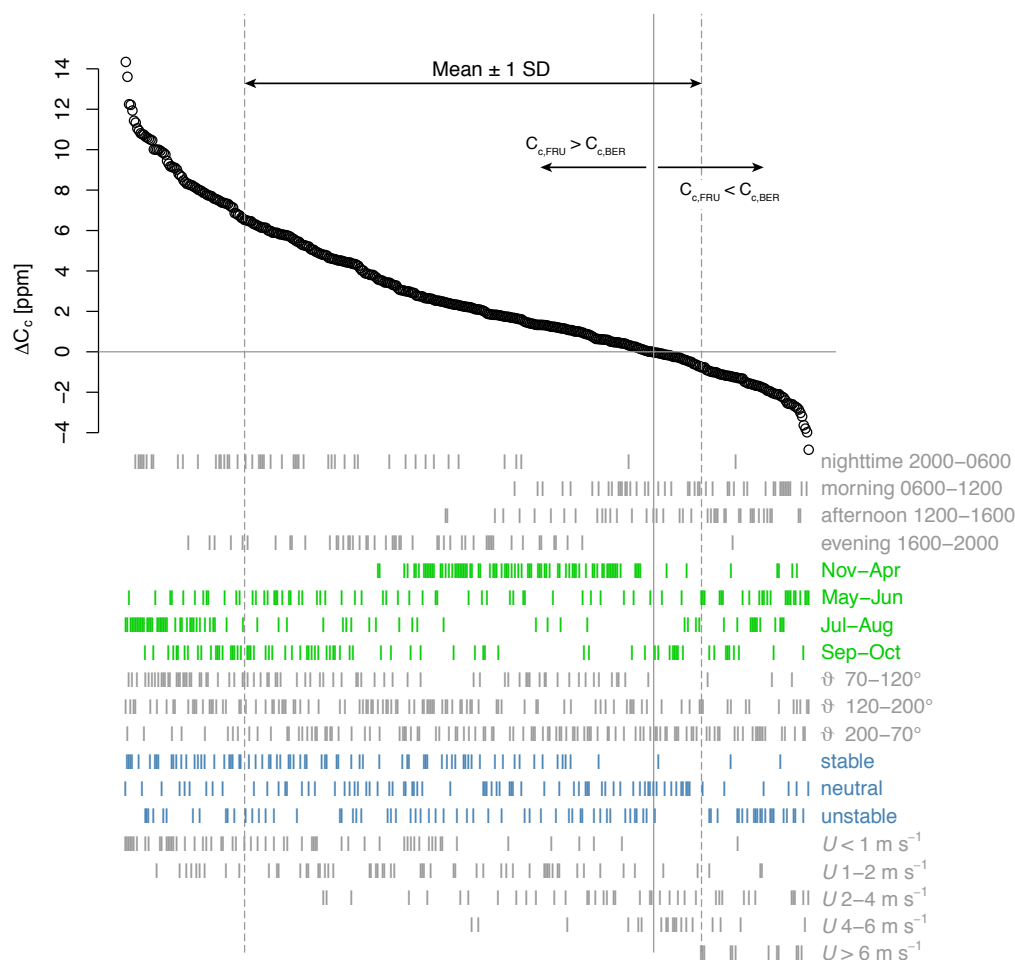


Fig. S5 Group means of differences in CO₂ mixing ratios measured at the FRU site in comparison with the BER site, sorted in descending order. Only group comparisons with statistically significant nonzero differences ($p < 0.05$) are shown. The mean difference \pm one standard deviation (i.e., 2.9 ± 3.7 ppm) is shown with two vertical dashed lines. The zero difference is shown with a horizontal and a vertical solid thin line. The vertical symbols below the graph indicate which class combination was used for the respective group mean

$< 1 \text{ m s}^{-1}$, and ϑ 120–200°), and the most negative value was -4.8 ppm (May–June, neutral z/L , U 2–4 m s^{-1} , and ϑ 120–200°).

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